



Royal Netherlands Institute for Sea Research

This is a pre-copyedited, author-produced version of an article accepted for publication, following peer review.

Balzano, S.; Jamieson, T. & Leterme, S. (2021). Changes in microbial communities during seawater pre-treatment within a desalination plant. *Aquatic Microbial Ecology*, 86: 63-68

Published version: <https://doi.org/10.3354/ame01958>

NIOZ Repository: <http://imis.nioz.nl/imis.php?module=ref&refid=335080>

[Article begins on next page]

The NIOZ Repository gives free access to the digital collection of the work of the Royal Netherlands Institute for Sea Research. This archive is managed according to the principles of the [Open Access Movement](#), and the [Open Archive Initiative](#). Each publication should be cited to its original source - please use the reference as presented.

When using parts of, or whole publications in your own work, permission from the author(s) or copyright holder(s) is always needed.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17

Changes in microbial communities during seawater pre-treatment within a desalination plant.

Sergio Balzano^{1,2*}, Tamar Jamieson^{3,4}, Sophie Leterme^{3,4}

¹ Stazione Zoologica di Napoli Anton Dohrn. Villa Comunale, 80121 Naples (Italy)

² Department of Marine Microbiology and Biogeochemistry, Netherland Institute for Sea Research (NIOZ), Landsdiep 4, Texel (Netherlands)

³ College of Science and Engineering, Flinders University, Sturt Road, Bedford Park, Adelaide, SA 5042 (Australia)

⁴ Institute for Nanoscale Science & Technology, Flinders University, Sturt Road, Bedford Park, Adelaide, SA 5042 (Australia)

*Corresponding author: Sergio.balzano@szn.it

18 **Running title:** microbes across seawater pre-treatment

19

20 **Abstract**

21 We analysed prokaryotic and eukaryotic communities across the seawater pre-treatment
22 system of a desalination plant, using 16S and 18S rRNA gene sequencing. The richness of
23 operational taxonomic units (OTUs) increased downstream of the pre-treatment system (RO
24 feedwater) compared to raw seawater for Archaea while it decreased for bacteria and protists.
25 Overall the RO feedwater was found to be enriched in ammonia-oxidising bacteria and
26 Archaea compared to raw seawater and also contained greater proportions of taxa typically
27 observed in aquatic biofilms and/or within other water treatment systems. Although, the
28 microbial load is reduced by the pre-treatment system, the increase in proportion of biofilm-
29 associated microbes suggest the presence of active microbial communities within multimedia
30 filters and other parts of the pre-treatment system that might increase biofouling risks.

31

32 **Keywords:** 16S rDNA, 18S rDNA, ammonium oxidation, desalination

33

34 Climate change and increase in human population worldwide are likely to exacerbate
35 freshwater scarcity issues in several regions (Naumann et al. 2018), highlighting the need for
36 alternative freshwater sources such as desalination. Most desalination plants use reverse
37 osmosis (RO) membranes for the conversion of seawater into freshwater and brine. To
38 remove most microbes and particles from the stream, seawater is usually pre-treated by
39 filtration and UV irradiation prior to RO. Although the microbial load decreases sharply after
40 seawater filtration, some of the microbes present after seawater pre-treatment are likely to
41 form biofilms on the RO membranes (Manes et al. 2011). For example, Bacteria are typically
42 enriched in Proteobacteria (Manes et al. 2011, Levi et al. 2016), whereas pennate diatoms and
43 other elongated species prevail among eukaryotes (Balzano et al. 2014) after pre-treatment
44 for SWRO. While bacterial communities have been widely characterised little is known on
45 Archaea and protists. Here, we assessed the composition of both prokaryotic and eukaryotic
46 communities occurring upstream (raw seawater) and downstream (RO feedwater) of a pre-
47 treatment system for SWRO by sequencing the V4 fragment of both the 16S and the 18S
48 rRNA genes.

49 Seawater was collected from the Penneshaw SWRO desalination plant, located on the
50 North-Eastern coast of Kangaroo Island, South Australia (Balzano et al. 2014). The pre-
51 treatment system consists in a medium pressure-ultra violet (MP-UV) disinfection unit,
52 multimedia and cartridge filters (Balzano et al. 2014). Nutrients were analysed every second
53 week over one-year period (18/07/2012 to 21/07/2013) from raw seawater, downstream of
54 the MP-UV treatment unit (Site 2), the multimedia filters (Site 3), the 15 µm cartridge filters
55 (Site 4), and from the RO feedwater, as described previously (Balzano et al. 2015b).
56 Molecular analyses were carried out on raw seawater and RO feedwater only, and samples
57 were collected 5 times over a 13-month period (Oct 2012, Dec 2012, Mar 2013, Jul 2013,
58 Nov 2013). One hundred-twenty L seawater were concentrated down to 2 L by tangential

59 flow filtration (Marie et al. 2010) and the concentrated sample was first pre-filtered using 10
60 μm cellulose filters and then filtered through 0.22-mm pore size Sterivex units (Millipore,
61 Billerica, MA). Cells were removed from the Sterivex units, the DNA extracted, the 18S
62 rRNA gene amplified, and both 18S and 16S rRNA genes sequenced using IonTorrent PGM
63 as described previously (Balzano et al. 2015a), whereas the V4 region of the 16S rRNA gene
64 was amplified using slight modifications (Supplementary Table S1) of the universal
65 prokaryotic primers V341F and 805R (Bowman et al. 2012) and PCR reactions consisted of
66 an initial denaturation at 98°C for 1 min, 30 cycles of 40 s at 98°C, 40 s at 53°C and 1 min at
67 72°C, and a final extension at 72°C for 1 min. Raw sequencing data were processed using the
68 python pipeline Quantitative Insight into Microbial Ecology (QIIME) (Caporaso et al. 2010);
69 read were filtered, clustered into operational taxonomic units (OTUs), and reads from
70 different samples were compared as described previously (Balzano et al. 2015a).

71 Raw seawater temperature ranged from 14.3 °C (July 2013) to 20.8 °C (March 2013) and
72 the salinity was stable around 36 psu (Balzano et al. 2015b). The abundance of both bacteria
73 and phytoplankton dropped dramatically across the multimedia filter, from Site 2 to Site 3
74 (Balzano et al. 2014). The concentration of ammonium dropped by half across the
75 multimedia filters (Fig. 1A), whereas the median concentration of NO_x increased from 0.35 to
76 1.24 μM (Fig. 1B) and both phosphate and silica did not change significantly (data not
77 shown). The decrease in the concentration of ammonium and the increase in NO_x , suggest
78 that nitrification was taking place within the multimedia filter. The lower ammonium to
79 nitrate ratios measured downstream the multimedia filters are likely to partially limit
80 microbial growth: heterotrophic bacteria in the water column are known to preferentially
81 uptake ammonium over nitrate as nitrogen source (Middelburg and Nieuwenhuize 2000,
82 Kumar et al. 2018).

83 Overall, we sampled a good portion of the microbial community (Supplementary Figure
84 1). OTU richness and diversity indices were generally lower in raw seawater than in RO
85 feedwater for Archaea, and higher for bacteria and protists (Table 1). Most taxonomical
86 changes across the pre-treatment plant were observed for Archaea compared to bacteria and
87 protists. The proportions of archaeal reads over the total 16S rRNA gene libraries were
88 significantly lower in raw seawater (0.5 to 3%) compared to RO feedwater (5 to 40%, Fig.
89 1C-D). Euryarchaeota, that typically dominate surface seawaters (Yin et al. 2013, Zhou et al.
90 2018), accounted for a large proportion of the genetic libraries of raw seawater, and their
91 contribution dropped dramatically in RO feedwater. In contrast, Nanoarchaeota and
92 *Nitrosoarchaeum* spp. (Thaumarchaeota) dominated the archaeal community in RO
93 feedwater being represented by 26% and 41% of archaeal reads, respectively (Fig. 1D.
94 *Nitrosoarchaeum* spp. are known to oxidise ammonium to nitrite (Konneke et al. 2005,
95 Pitcher et al. 2011) and the increase in the abundance of reads associated with this genus in
96 RO feedwater (Fig. 1C-D), suggest a potential role of *Nitrosoarchaeum* spp. in the
97 ammonium oxidation observed here (Fig. 1A-B). *Nitrosoarchaeum* spp. were previously
98 sequenced in multimedia filters of a drinking water treatment system in which ammonium
99 oxidation was also found to occur (Bai et al. 2013).

100 Bacteria were dominated by α -Proteobacteria with high contributions from Bacteroidetes,
101 Synechococcales, and γ -Proteobacteria. Melainobacteria, γ -Proteobacteria, and
102 Verrucomicrobia tended to be more represented in genetic libraries from the RO feedwater.
103 Protists were mostly represented by dinoflagellates, ciliates, Syndiniales, Stramenopiles,
104 Rhizaria, Archaeplastida, and Opisthokonta, with ciliates, Rhizaria, and Opisthokonta being
105 more represented in RO feedwater (Supplementary Figure S2, Supplementary Tables S2-S3).

106 ANOSIM analyses revealed significant differences in Bray-Curtis dissimilarities as well as
107 unweighted and weighted UniFrac distances, between raw seawater and RO feedwater

108 communities for Archaea, bacteria and, to a lesser extent, protists. In contrast, microbial
109 communities sampled at different dates did not show significant differences (data not shown).
110 This indicates that microbial community differences across the SWRO pre-treatment system
111 were greater than seasonal differences on the same sampling site. Furthermore, taxonomic
112 differences across Penneshaw seawater pre-treatment system were greater for prokaryotes
113 compared to eukaryotes.

114 Most archaeal OTUs (58%) were only detected in RO feedwater whereas this proportion
115 was lower for bacteria (26%), and protists (26%) (Supplementary Figure S3). Data suggest
116 that at least some of the archaeal and bacterial taxa sequenced here were likely to stably
117 persist within Penneshaw SWRO plant during different seasons. Overall, 17 archaeal OTUs,
118 51 bacterial OTUs, and 6 eukaryotic OTUs were found occur in all RO feedwater samples
119 and their contribution to the overall community was significantly (p -value > 0.01) higher than
120 that observed in raw seawater (Fig. 2, Supplementary Table S4). RO feedwater persistent and
121 enriched OTUs are represented by ammonium oxidising microbes, and biofilm-associated
122 taxa typically observed in soil, sediment or different water treatment plants.

123 Ammonium oxidising microbes enriched in RO feedwater include the Archaea
124 *Nitrosoarchaeum* and *Nitrosopumilus* (Fig. 2) that were previously detected within pre-
125 treatment systems of SWRO desalination plants (Hong et al. 2016, Jeong et al. 2016) and the
126 bacterium *Nitrospira* sp., which was found in wastewater treatment plants (Keuter et al.
127 2011) and biofilters of recirculating aquaculture systems (Brown et al. 2013). Current data
128 thus suggest that *Nitrosoarchaeum*, *Nitrosopumilus*, and *Nitrospira* representatives found
129 here were likely to colonise the multimedia filters of the seawater pre-treatment being
130 responsible for the ammonium oxidation measured (Fig. 1). Other OTUs that are significantly
131 enriched in RO feedwater mostly belong to taxa that have been previously found in RO
132 membranes or RO feedwater such as *Pseudoalteromonas* spp. and Cryomorphaceae (Chun et

133 al. 2012, Nagaraj et al. 2019) or biofilm forming taxa such as Melainabacteria (Rehman et al.
134 2020) and Hartmannulidae ciliates (Xu et al. 2014).

135 Taxa persistent in RO feedwater are likely to either occur in South Australian coastal
136 waters throughout the year and systematically passing through the pre-treatment system
137 because of some specific features (size, shape), or to be part of a persistent community
138 present within Penneshaw pre-treatment system. While some of these taxa, especially
139 ammonium oxidizers, are likely to derive from multimedia filters, others might be associated
140 with biofilms present in other surfaces of the pre-treatment system. It has been demonstrated
141 that biofilm-associated microbes present along pre-treatment systems can behave as microbial
142 reservoirs potentially enhancing the risks of RO membrane biofouling (Levi et al. 2016).

143 In spite of the UV treatment and the presence of several multimedia and cartridge filters,
144 Penneshaw SWRO plant harbours an RO feedwater-specific community which mostly
145 includes prokaryotic microbes. Some of these microbes are potentially involved in the
146 oxidation of ammonium to nitrite and nitrate (*Nitrosoarchaeum* spp., *Nitrosopumilus* spp.,
147 and *Nitrospira* spp.) within the multimedia filters. Most RO feedwater-specific microbes
148 were previously isolated or sequenced from different water treatment facilities and some of
149 them can potentially cause biofouling on the RO membrane. In contrast with other
150 desalination plants, chemical disinfection is not applied in Penneshaw SWRO plant to
151 decrease the environmental impact, thus potentially leading to an increased microbial load in
152 RO feedwater. Microbial communities similar to the RO feedwater-specific community
153 found here are likely to occur in other disinfection-free SWRO plants. Our results thus
154 provide insights on the bacteria and Archaea potentially causing biofouling and can
155 contribute, to future research, to design effective strategies to minimise biofouling.

156

157 **Acknowledgements**

158 The authors acknowledge the financial support of the National Centre of Excellence in
159 Desalination Australia (NCEDA) which is funded by the Australian Government
160 through the National Urban Water and Desalination Plan. The authors are also grateful to P.
161 Meacham, T. Kirby, and N. Nedelkov for their assistance in sampling at the Penneshaw
162 desalination plant. We also thank Xavier Denis, Camille Moreau and Jan-Georg Jendyk for
163 helping with sampling and processing nutrient samples.

164

References

- 166 Bai, Y. H., R. P. Liu, J. S. Liang and J. H. Qu (2013). "Integrated Metagenomic and Physiochemical
167 Analyses to Evaluate the Potential Role of Microbes in the Sand Filter of a Drinking Water
168 Treatment System." *Plos One* **8**(4).
- 169 Balzano, S., E. Abs and S. C. Leterme (2015a). "Protist diversity along a salinity gradient in a coastal
170 lagoon." *Aquatic Microbial Ecology* **74**(3): 263-277.
- 171 Balzano, S., A. V. Ellis, C. Le Lan and S. C. Leterme (2015b). "Seasonal changes in phytoplankton on
172 the north-eastern shelf of Kangaroo Island (South Australia) in 2012 and 2013." *Oceanologia*
173 **57**(3): 251-262.
- 174 Balzano, S., C. Le Lan, A. V. Ellis, H. Compas, K. Newton, T. Jamieson, M. Brown and S. C. Leterme
175 (2014). "Evaluation of transparent exopolymer particles and microbial communities found
176 post-UV light, multimedia and cartridge filtration pre-treatment in a SWRO plant."
177 *Desalination and Water Treatment* **56**(6): 1427-1439.
- 178 Bowman, J. S., S. Rasmussen, N. Blom, J. W. Deming, S. Rysgaard and T. Sicheritz-Ponten (2012).
179 "Microbial community structure of Arctic multiyear sea ice and surface seawater by 454
180 sequencing of the 16S RNA gene." *Isme Journal* **6**(1): 11-20.
- 181 Brown, M. N., A. Briones, J. Diana and L. Raskin (2013). "Ammonia-oxidizing archaea and nitrite-
182 oxidizing nitrospiras in the biofilter of a shrimp recirculating aquaculture system." *Fems*
183 *Microbiology Ecology* **83**(1): 17-25.
- 184 Caporaso, J. G., J. Kuczynski, J. Stombaugh, K. Bittinger, F. D. Bushman, E. K. Costello, N. Fierer, A. G.
185 Pena, J. K. Goodrich, J. I. Gordon, G. A. Huttley, S. T. Kelley, D. Knights, J. E. Koenig, R. E. Ley,
186 C. A. Lozupone, D. McDonald, B. D. Muegge, M. Pirrung, J. Reeder, J. R. Sevinsky, P. J.
187 Tumbaugh, W. A. Walters, J. Widmann, T. Yatsunencko, J. Zaneveld and R. Knight (2010).
188 "QIIME allows analysis of high-throughput community sequencing data." *Nature Methods*
189 **7**(5): 335-336.
- 190 Chun, Y., P. T. Ha, L. Powell, J. Lee, D. Kim, D. Choi, R. W. Lovitt, I. S. Kim, S. S. Mitra and I. S. Chang
191 (2012). "Exploring microbial communities and differences of cartridge filters (CFs) and
192 reverse osmosis (RO) membranes for seawater desalination processes." *Desalination* **298**:
193 85-92.
- 194 Hong, P. Y., N. Moosa and J. Mink (2016). "Dynamics of microbial communities in an integrated
195 ultrafiltration-reverse osmosis desalination pilot plant located at the Arabian Gulf."
196 *Desalination and Water Treatment* **57**(35): 16310-16323.
- 197 Jeong, S., K. Cho, H. Bae, P. Keshvardoust, S. A. Rice, S. Vigneswaran, S. Lee and T. Leiknes (2016).
198 "Effect of microbial community structure on organic removal and biofouling in membrane
199 adsorption bioreactor used in seawater pretreatment." *Chemical Engineering Journal* **294**:
200 30-39.
- 201 Keuter, S., M. Kruse, A. Lipski and E. Spieck (2011). "Relevance of Nitrospira for nitrite oxidation in a
202 marine recirculation aquaculture system and physiological features of a Nitrospira marina-
203 like isolate." *Environmental Microbiology* **13**(9): 2536-2547.
- 204 Konneke, M., A. E. Bernhard, J. R. de la Torre, C. B. Walker, J. B. Waterbury and D. A. Stahl (2005).
205 "Isolation of an autotrophic ammonia-oxidizing marine archaeon." *Nature* **437**(7058): 543-
206 546.
- 207 Kumar, S., P. S. Bhavya, R. Ramesh, G. V. M. Gupta, F. Chiriboga, A. Singh, I. Karunasagar, A. Rai, A.-S.
208 Rehnstam-Holm, L. Edler and A. Godhe (2018). "Nitrogen uptake potential under different
209 temperature-salinity conditions: Implications for nitrogen cycling under climate change
210 scenarios." *Marine Environmental Research* **141**: 196-204.
- 211 Levi, A., E. Bar-Zeev, H. Elifantz, T. Berman and I. Berman-Frank (2016). "Characterization of
212 microbial communities in water and biofilms along a large scale SWRO desalination facility:
213 Site-specific prerequisite for biofouling treatments." *Desalination* **378**: 44-52.

214 Manes, C. L. D., N. West, S. Rapenne and P. Lebaron (2011). "Dynamic bacterial communities on
215 reverse-osmosis membranes in a full-scale desalination plant." *Biofouling* **27**(1): 47-58.
216 Marie, D., X. L. Shi, F. Rigaut-Jalabert and D. Vaultot (2010). "Use of flow cytometric sorting to better
217 assess the diversity of small photosynthetic eukaryotes in the English Channel." *Fems*
218 *Microbiology Ecology* **72**(2): 165-178.
219 Middelburg, J. J. and J. Nieuwenhuize (2000). "Nitrogen uptake by heterotrophic bacteria and
220 phytoplankton in the nitrate-rich Thames estuary." *Marine Ecology Progress Series* **203**: 13 -
221 21.
222 Nagaraj, V., L. Skillman, D. Li, Z. Xie and G. Ho (2019). "Culturable bacteria from a full-scale
223 desalination plant: Identification methods, bacterial diversity and selection of models based
224 on membrane-biofilm community." *Desalination* **457**: 103-114.
225 Pitcher, A., E. C. Hopmans, A. C. Mosier, S. J. Park, S. K. Rhee, C. A. Francis, S. Schouten and J. S. S.
226 Damste (2011). "Core and Intact Polar Glycerol Dibiphytanyl Glycerol Tetraether Lipids of
227 Ammonia-Oxidizing Archaea Enriched from Marine and Estuarine Sediments." *Applied and*
228 *Environmental Microbiology* **77**(10): 3468-3477.
229 Rehman, Z. U., L. Fortunato, T. Y. Cheng and T. Leiknes (2020). "Metagenomic analysis of sludge and
230 early-stage biofilm communities of a submerged membrane bioreactor." *Science of the Total*
231 *Environment* **701**: 10.
232 Yin, Q., B. B. Fu, B. Y. Li, X. C. Shi, F. Inagaki and X. H. Zhang (2013). "Spatial Variations in Microbial
233 Community Composition in Surface Seawater from the Ultra-Oligotrophic Center to Rim of
234 the South Pacific Gyre." *Plos One* **8**(2): 12.
235 Xu, H. L., W. Zhang, Y. Jiang and E. J. Yang (2014). "Use of biofilm-dwelling ciliate communities to
236 determine environmental quality status of coastal waters." *Science of the Total Environment*
237 **470**: 511-518.
238 Zhou, J., X. Song, C. Y. Zhang, G. F. Chen, Y. M. Lao, H. Jin and Z. H. Cai (2018). "Distribution Patterns
239 of Microbial Community Structure Along a 7000-Mile Latitudinal Transect from the
240 Mediterranean Sea Across the Atlantic Ocean to the Brazilian Coastal Sea." *Microbial Ecology*
241 **76**(3): 592-609.

242

243

244

245

246 **Figure Legend**

247

248

249 **Figure 1.** Box and whisker plots highlighting changes in the concentration of (A)
250 ammonium, as well as (B) nitrate and nitrite within the seawater pre-treatment system of
251 Penneshaw desalination plant. Each box includes 25th and 75th percentiles from each
252 parameter and lines within the boxes represent median values, whereas solid lines outside the
253 boxes correspond to the 10th and the 90th percentiles. Taxonomic composition (left axis) and
254 total contribution to the total 16S rRNA gene libraries (right axis) of the Archaea sequenced
255 from (C) raw seawater and (D) RO feedwater in the seawater pre-treatment system of
256 Penneshaw SWRO plant inferred from 16S rRNA gene sequencing.

257

258

259 **Figure 2.** Distribution of all the OTUs found to be persistent in RO feedwater and enriched
260 (t-test, p-value > 0.01) compared to raw seawater and recovered from RO feedwater at all
261 time points (persistent community). The taxonomic affiliation of each OTU is indicated next
262 to the OTU code.

263

264 **Table 1.** Sequencing results and analysis of microbial diversity. OTUs: operational
 265 taxonomic units
 266

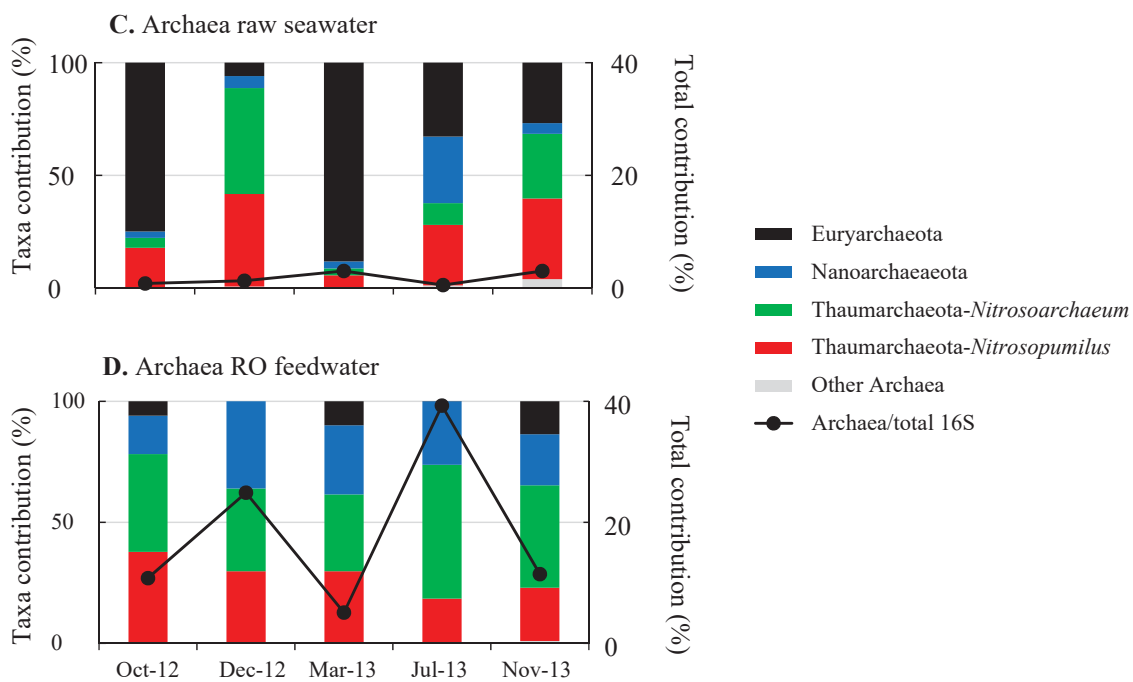
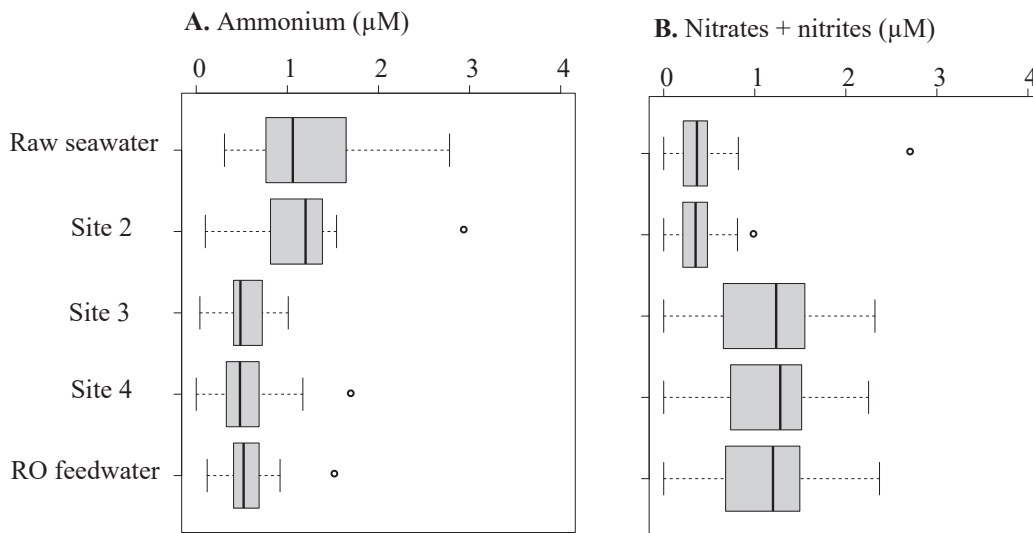
	Sampling site	Sampling date	No of reads	No of OTUs
Archaeal 16S	Raw seawater	24/10/12	117	30
		19/12/12	168	26
		27/03/13	385	40
		03/07/13	56	25
		27/11/13	387	89
		Total	1,113	136
	RO feedwater	24/10/12	1424	74
		19/12/12	3231	104
		27/03/13	674	86
		03/07/13	5252	124
		27/11/13	1530	133
	Total	12,111	260	
Bacterial 16S	Raw seawater	24/10/12	9390	1036
		19/12/12	12098	1191
		27/03/13	9891	1283
		03/07/13	10071	2047
		27/11/13	9586	1983
		Total	51,036	4449
	RO feedwater	24/10/12	10079	1041
		19/12/12	10038	1098
		27/03/13	11928	952
		03/07/13	7947	1192
		27/11/13	10608	1226
	Total	50,600	3089	
Total				
Protists	Raw seawater	24/10/12	17,242	1886
		19/12/12	17,242	919
		27/03/13	17,242	1507
		03/07/13	17,242	1650
		27/11/13	17,242	2499
		Total	86,210	5781
	RO feedwater	24/10/12	17,242	963
		19/12/12	17,242	410
		27/03/13	17,242	992
		03/07/13	17,242	587
		27/11/13	17,242	1907
	Total	86,210	3771	

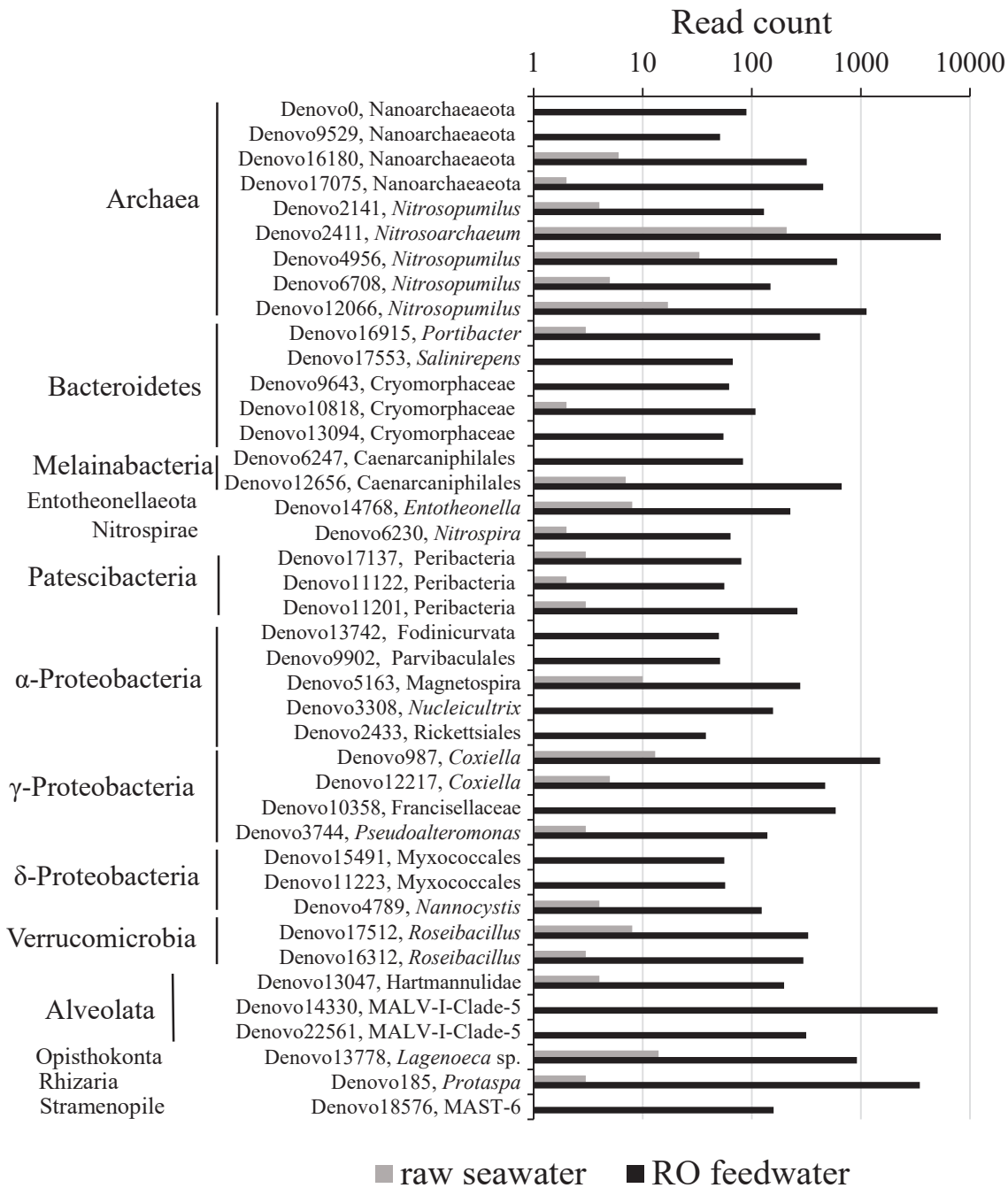
267
 268

269 **Table 2.** Analysis of similarity (ANOSIM) highlighting differences between microbial
 270 communities sampled from raw seawater and RO feed water^a
 271

Community	Bray Curtis		Unweighted Unifrac		Weighted Unifrac	
	test	p-value	test	p-value	test	p-value
Archaea (16S)	<u>0.44</u>	0.02	0.89	0.008	0.45	0.007
Bacteria (16S)	0.4	0.005	<u>1</u>	0.01	<u>0.31</u>	0.036
Plastidic protists (16S)	<u>0.35</u>	0.029	0.1	0.25	0.21	0.13
All protists (18S)	<u>0.34</u>	0.012	<u>0.41</u>	0.02	<u>0.42</u>	0.011

272
 273 ^a Significant (p-value < 0.05) and highly significant (p-value < 0.01) correlations are underlined and in bold,
 274 respectively.





Supplementary data

Supplementary Figure S1. Accumulation curves, representing the OTU richness in each sample, as function of the number of reads sequenced for (A-B) the 16S and (C-D) the 18S rRNA gene libraries in (A, C) raw seawater and (B, D) RO feedwater.

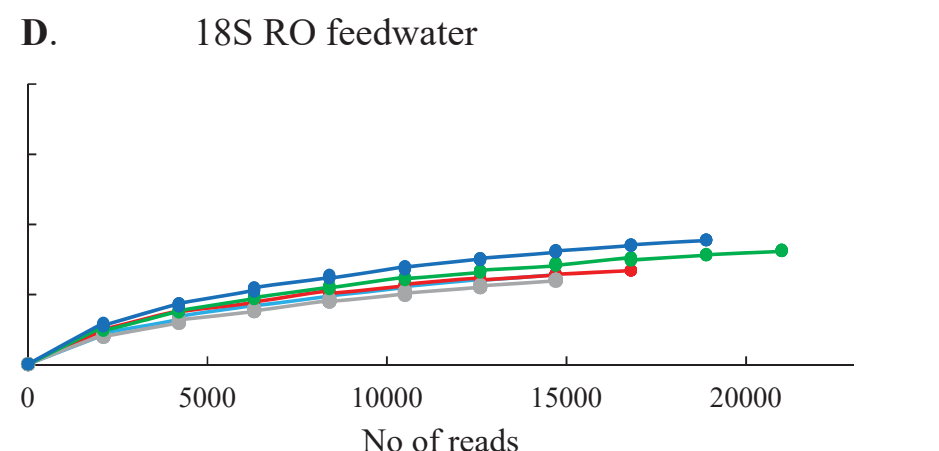
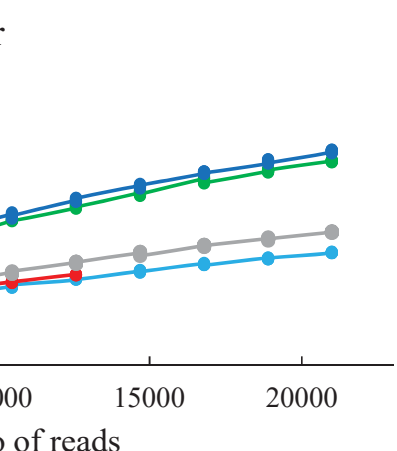
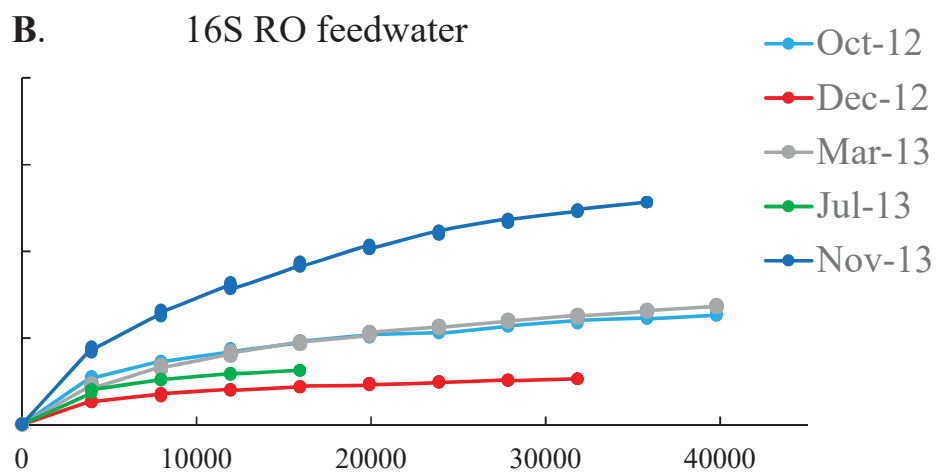
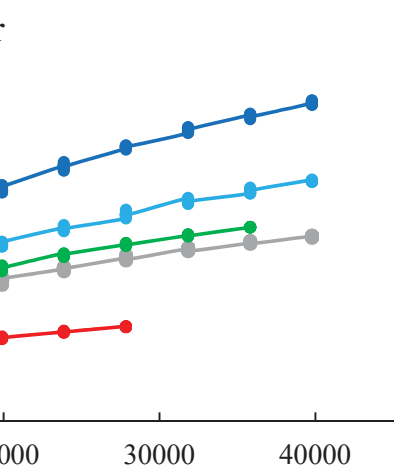
Supplementary Figure S2. Venn diagrams representing similarities and differences between the microbial communities of (A) Archaea, (B) bacteria, (C) plastidic protists, and (D) total protists, identified in raw seawater (light blue circles) and RO feed water (pink circles) of the Penneshaw SWRO plant. Values indicate the number of OTUs followed by their percentage contribution to total reads. The two small circles within each pink circle represent the OTUs found to be persistent (occurring at each time point) in RO feed water as well as in all samples (both raw seawater and RO feedwater), respectively.

Supplementary Table S1. Details of primers, adapters and barcodes used in the present study for Ion Torrent sequencing

Supplementary Table S2. Taxonomic affiliation and distribution within the different samples for the OTUs (97 % identity) found here within the prokaryotic 16S rRNA gene dataset.

Supplementary Table S3. Taxonomic affiliation and distribution within the different samples for the OTUs (97 % identity) found here within the 18S dataset

Supplementary Table S4. Statistics of the OTUs from the 16S rRNA gene that were found to be significantly more abundant in RO feedwater compared to raw seawater for the 16S rRNA gene



representing the OTU richness in each sample, as function of the number of reads sequenced for (A-B) the 16S and (C-D) the 18S rDNA in (A, C) raw seawater and (B, D) RO feedwater.