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Impact of riverine suspended particulate matter on the branched
glycerol dialkyl glycerol tetraether composition of lakes: The outflow
of the Selenga River in Lake Baikal (Russia)

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1 **ABSTRACT**

2 Branched glycerol dialkyl glycerol tetraethers (brGDGTs) are bacterial membrane lipids
3 occurring in several environments, including soils, rivers and lakes, whose distribution varies with
4 temperature and pH, although this dependence is apparently not the same for the different producing
5 environments. Mixing of brGDGT sources may thus complicate palaeoenvironmental reconstruction.
6 The extent to which brGDGTs in a lake outflow reflect the brGDGT distribution delivered by
7 upstream rivers was studied for Lake Baikal (Russia), one of the largest freshwater lakes worldwide.
8 Fifteen brGDGTs were quantified in suspended particulate matter (SPM) of the Selenga River and its
9 outflow in the lake. The river and lake SPM had rather different brGDGT distributions. The riverine
10 brGDGT distribution was still apparent in the SPM of the lake surface water 5 km from the river
11 mouth, but shifts in the brGDGT distribution were already apparent in the SPM of the surface water
12 after 1 km. Based on the brGDGT distributions of the SPM of the Selenga outflow and that of the lake,
13 conservative mixing between the river and the lake brGDGT distributions could not fully explain the
14 observed shifts in brGDGT distributions. Both preferential degradation and in-situ production of
15 brGDGTs in the surface and, especially, bottom water of the river outflow were potentially
16 responsible. This implies that a riverine lipid distribution delivered to a lake can be modified prior to
17 being transported downstream. The lacustrine brGDGT distribution, that possibly reflected a mixture
18 of mountainous and Selenga River SPM, was not recognized in downstream Yenisei River SPM. The
19 watershed of Lake Baikal thus does not seem to contribute to the brGDGTs transported to the marine
20 system. As many large rivers have major lakes in their watershed, this has implications for
21 palaeoclimate reconstruction from river fan sediments globally.

22

23

24 **KEYWORDS**

25 brGDGTs, 6-methyl, in-situ production, degradation, Selenga River, Lake Baikal

26 **1. Introduction**

27 Branched glycerol dialkyl glycerol tetraethers (brGDGTs) are bacterial membrane lipids found
28 in a variety of settings: soils, lacustrine and marine suspended particulate matter (SPM) and sediments
29 and hotspots. Their source organisms probably fall within the Acidobacteria, based on
30 environmental (Weijers et al., 2009; Peterse et al., 2010) and culture studies (Sinninghe Damsté et al.,
31 2011; 2014). Their main application is as palaeoclimate proxies. In a dataset of global soils, the
32 structural diversity of the brGDGT components was shown to correlate with the prevailing soil pH and
33 mean annual air temperature (MAAT; Weijers et al., 2007a). Nine brGDGT components were
34 described that possess 4 to 6 methyl substituents (branches) on the linear C₂₈ alkyl chains (Sinninghe
35 Damsté et al., 2000) and contain up to two cyclopentyl moieties formed by internal cyclization (Fig. 1;
36 Schouten et al., 2000; Weijers et al., 2006). The cyclization of branched tetraethers (CBT) and
37 methylation of branched tetraethers (MBT) are two brGDGT indices (Weijers et al., 2007a) that have
38 been successfully applied to reconstruct the palaeoclimatic changes in palaeosoils (e.g. Peterse et al.,
39 2011), speleothems (Blyth and Schouten, 2013), lake sediments (e.g. Niemann et al., 2012), but
40 initially in marine sediments (e.g. Weijers et al., 2007b; Bendle et al., 2010). Prior to their
41 incorporation into marine and lacustrine sediments, they were understood to be eroded from soils and
42 transported by rivers through upstream lakes. The brGDGT distribution was thus assumed to reflect
43 the distribution present in watershed soils and to remain unaltered during this process.

44 Contrasting brGDGT distributions between rivers, lakes and their surrounding soils provided
45 the first indications for in-situ production of brGDGTs in freshwater aquatic systems (e.g. Tierney and
46 Russell, 2009). Although the temperature dependence is different from that in soils, the distribution of
47 aquatic brGDGTs in lakes also varies with prevailing MAAT, and to some extent with depth and pH
48 (e.g. Tierney et al., 2010; Pearson et al., 2011; Sun et al., 2011; Loomis et al., 2012, 2014a). There is
49 thus a growing body of evidence supporting in-situ production of brGDGTs in lakes, but the niche of
50 the source organism has not been constrained. The concentration of lacustrine brGDGTs was shown to
51 increase below the lake thermocline, pointing towards a preference for environments with low O₂
52 concentration (e.g. Sinninghe Damsté et al., 2009; Buckles et al., 2014a). However, a recent study of
53 brGDGTs in a temperate lake (Loomis et al., 2014b) found that brGDGTs were produced throughout

54 the water column. Further possible mechanisms influencing the brGDGT distributions are shifts in
55 bacterial community, possibly prompted by a large shift in nutrients or by the transition between river
56 and lake biomes (Loomis et al., 2014b and references therein). Although in-situ production of
57 brGDGTs has been described to occur in rivers (e.g. Kim et al., 2012; Zell et al., 2013, 2014; De
58 Jonge et al., 2014a), a lacustrine in-situ produced distribution may be significantly different from that
59 of its inflowing river (Tierney and Russell, 2009; Buckles et al., 2014b). As large lakes are often
60 present in large river drainage basins, lacustrine in-situ production may result in the introduction of
61 lacustrine brGDGTs in downstream rivers and ultimately in marine sediments. Large lakes in the
62 drainage basin of river systems may thus have an effect on palaeoclimate brGDGT reconstruction for
63 river fan marine sediments.

64 The aim of this study was to investigate the extent to which the brGDGT distribution delivered
65 by a river can propagate in a large lake and to compare this effect with the brGDGT distribution
66 exported from the lake by river outflow. Furthermore, we tried to constrain the environmental
67 parameters that influence lacustrine in-situ production of brGDGTs. Although in-situ production of
68 brGDGTs in lakes has been extensively documented (e.g. Tierney and Russell, 2009; Loomis et al.,
69 2011, 2014b; Buckles et al., 2014a), this is the first study to evaluate the delivery and export of
70 riverine brGDGTs to and from a lake system. Furthermore, the above previous studies are all based on
71 a dataset of nine brGDGTs, as the analytical procedure used did not allow separating the recently
72 described 6-Me brGDGTs (De Jonge et al., 2013). The abundance of these novel brGDGTs
73 was recently shown to be high in a Siberian River system (De Jonge et al., 2014a) and to be highly
74 variable in a set of globally distributed soils (De Jonge et al., 2014b).

75 This study describes the full suite of fifteen brGDGTs in a major river (Selenga River) that
76 drains northern Mongolia and southern Siberia before and after its inflow to the world's largest
77 freshwater lake (Lake Baikal). We evaluate the brGDGT concentration and distribution in the Selenga
78 River outflow, where in-situ production and preferential degradation could possibly affect the
79 lacustrine brGDGT distributions. Furthermore, the distribution exported from the lake was compared
80 with the brGDGT distribution in both Selenga River and the mountainous Irkut River, to evaluate
81 whether or not riverine brGDGTs alone could explain the lacustrine brGDGT signature exported.

82

83 **2. Geographical setting**

84 The Selenga River originates in the mountainous parts of Mongolia, draining large parts of
85 Mongolia and southern Siberia (442 000 km²; Fig. 2b). It is the main tributary of Lake Baikal, with a
86 drainage area 82% of the total drainage area of the lake. It transports 57.8 km³/yr, which accounts for
87 ca. 50% (Votintsev et al., 1985) of the total water input to the lake. Furthermore, it contributes ca.
88 80% of the total suspended solids delivered by the tributaries to the lake (Votintsev et al., 1985). The
89 other tributaries (Fig. 2b) drain the steep, mountainous watershed that borders the lake on the North
90 and the East. The lake is one of the largest in the world, containing ca. 20% of the Earth's fresh liquid
91 surface water (23,000 km³). The lake water is well mixed and aerated. The chemical and biological
92 parameters of the Selenga River outflow have been studied by Maksimenko et al. (2008) and
93 Sorokovikova et al. (2012). The mixing zone was described to be biologically very active, fueled by
94 the riverine NO₃⁻ and PO₄³⁻. Based on changes in water chemistry and the phyto- and bacterioplankton,
95 the latter authors concluded that, during the summer months, the mixing zone extends from 1–3 km
96 downstream off the Selenga River mouth, after which a lake signature dominates. Maksimenko et al.
97 (2008) studied the microbial communities of the Selenga River outflow and concluded that a fully
98 lacustrine composition occurs between 5 and 7 km downstream of the river mouth.

99

100 **3. Material and methods**

101 *3.1. Collection of suspended particulate matter (SPM)*

102 Table 1 lists the SPM samples investigated and the locations of the sampling stations are shown
103 in Fig. 2. In July 2010, 100 l of Lake Baikal water was sampled from the shoreline at the point where
104 it drains into the Angara River (Ba), further referred to as 'Lake Baikal outflow'). Between 10-20 L of
105 the Selenga River outflow into Lake Baikal (SRM, S1, S3, S5, B1, B3, B5) were sampled later in July
106 2010 (Fig. 2a). It was sampled at the surface at the river mouth (Harauz tributary; SRM), and surface
107 and bottom waters were sampled at 1 (S1 and B1), 3 (S3 and B3) and 5 km (S5 and B5) from the river
108 mouth. The remaining samples were filtered using the same type of 0.7 µm GF/F glass fiber filters.

109 The bulk parameters of the riverine sites Y1, Y2, MIR and SR (Fig. 2) have been discussed by De
110 Jonge et al., 2014a and the GDGT contents have been described by De Jonge et al. (submitted).

111

112 *3.2. Lipid extraction and analysis*

113 The freeze-dried filters were extracted using a modified Bligh and Dyer method as described by
114 Pitcher et al. (2009). The samples were ultrasonically extracted 3x for 10 min using a
115 MeOH/dichloromethane (DCM)/phosphate buffer 10:5:4 (v/v/v). The extract was separated into a core
116 lipid (CL) fraction and intact polar lipid (IPL) fraction over a small silica column, using a procedure
117 modified from Pitcher et al. (2009), using hexane/EtOAc (1:1, v/v) and MeOH as eluents. An aliquot
118 of the IPL fraction was analyzed directly for CLs to check for potential carryover into the IPL fraction.
119 In order to analyze IPLs as CLs, half of the extract was refluxed for a minimum of 2 h in 1.5 N HCl in
120 MeOH. However, IPL-derived brGDGTs were often below the limit of detection, indicating that the
121 amount of IPLs was insufficient for quantification, so we refrain from reporting the IPL brGDGTs
122 composition of Selenga River and Lake Baikal. All brGDGTs were quantified against a known
123 amount of C₄₆ GDGT standard (Huguet et al., 2006) added to the CL fraction before filtration through
124 a 0.45 µm PTFE filter and to the IPL fraction before the separation preceding the acid hydrolysis.

125 Fractions were analyzed using high performance liquid chromatography-atmospheric pressure
126 chemical ionization-mass spectrometry (HPLC-APCI-MS), as described by De Jonge et al. (2014a).
127 Detection was achieved in selected ion monitoring mode (SIM; Schouten et al., 2007) using *m/z* 744
128 for the internal standard, *m/z* 1292 for crenarchaeol and *m/z* 1050, 1048, 1046, 1036, 1034, 1032,
129 1022, 1020 and 1018 for brGDGTs. Agilent Chemstation software was used to integrate peak areas in
130 the mass chromatograms of the [M+H]⁺ ions.

131

132 *3.3. Calculation of GDGT-based ratios and proxies*

133 The isomer ratio (IR) represents the fractional abundance of the penta- and hexamethylated 6-
134 methyl (6-Me) brGDGTs vs. the total penta- and hexamethylated brGDGTs (modified after De Jonge
135 et al., 2014a):

$$136 \quad \text{IR} = (\text{IIa}', \text{b}', \text{c}' + \text{IIIa}', \text{b}', \text{c}') / (\text{IIa}, \text{b}, \text{c} + \text{IIIa}, \text{b}, \text{c} + \text{IIa}', \text{b}', \text{c}' + \text{IIIa}', \text{b}', \text{c}') \quad 1$$

137 The roman numerals refer to the GDGTs indicated in Fig. 1. The index includes the non-
138 cyclopentane containing and the cyclopentane containing components of brGDGTs II, III, II' and III'.

139 The branched and isoprenoid tetraether (BIT) index was calculated according to Hopmans et al.
140 (2004):

$$141 \text{ BIT index} = (Ia + IIa + IIIa + IIa' + IIIa') / (Ia + IIa + IIIa + IIa' + IIIa' + IV) \quad 2$$

142 Ia, IIa, IIIa, IIa' and IIIa' are brGDGTs and IV is the isoprenoid GDGT (iGDGT) crenarchaeol,
143 a source-specific GDGT for mesophilic Thaumarchaeota (Sinninghe Damsté et al., 2002; Pearson et
144 al., 2004).

145 We calculated a reconstructed pH using the modified cyclization of branched tetraether CBT'
146 index (De Jonge et al., 2014b):

$$147 \text{ CBT}' = {}^{10}\log[(Ic + IIa' + IIb' + IIc' + IIIa' + IIIb' + IIIc') / (Ia + IIa + IIIa)]. \quad 3$$

$$148 \text{ pH} = 7.15 + 1.59 \times \text{CBT}' \quad 4$$

149 Mean summer temperature (MST) is calculated as an approximation for the summer lake water
150 temperature recorded by aquatic brGDGT distributions. To this end, we applied the calibration
151 developed by Pearson et al. (2011):

$$152 \text{ MST (}^\circ\text{C)} = 20.9 + 89.1 \times [Ib] - 12 \times ([IIa] + [IIa']) - 20.5 \times [IIIa] \quad 5$$

153 Here, we explicitly state that the fractional abundances of IIa and IIa' have to be summed, as
154 these brGDGTs co-eluted under the chromatographic conditions used by Pearson et al. (2011).
155 However, conditions did allow separation of brGDGTs IIIa and IIIa', and only the first eluting one
156 (IIIa) was used in the calibration dataset (E. Pearson, personal communication). The square brackets in
157 Eq. 5 indicate that we used the fractional abundance, i.e. the value relative to the sum of all the
158 brGDGTs (Ia + Ib + Ic + IIa + IIa' + IIb + IIb' + IIc + IIc' + IIIa + IIIa' + IIIb + IIIb' + IIIc + IIIc').

159

160 *3.4. Environmental parameters and bulk geochemical analysis*

161 The pH and temperature of the river and lake water were measured immediately after sampling,
162 except for the sample SR, whose parameters were measured after transport. The particulate organic

163 carbon (POC) and $\delta^{13}\text{C}_{\text{POC}}$ content of the river and lacustrine SPM samples were measured on the filter
164 using a Flash 2000 Organic Elemental Analyzer.

165

166 *3.5. Numerical analysis*

167 Principal component analysis (PCA) based on the correlation matrix was performed using the R
168 software package for statistical computing. We performed an unconstrained Q-mode PCA on the
169 standardized relative brGDGT values in SPM from the Selenga River, Lake Baikal and mountainous
170 Irkut River, using 11 brGDGTs. The brGDGTs IIc, IIIc, IIc' and IIIc' were excluded from this
171 analysis, as they were absent from 4 to 7 sample sites. The brGDGT scores were calculated
172 proportional to the eigenvalues, and the site scores were calculated as the weighted sum of the species
173 scores. The environmental parameters (point measurement of water pH and temperature) were plotted
174 a posteriori in the resulting ordination space, in case a significant correlation ($p>0.05$) with the
175 principal components (PCs) was present.

176

177 **4. Results**

178 The brGDGT distribution of the SPM in the Selenga River was determined, before and after its
179 outflow into Lake Baikal. The river SPM was collected ca. 150 km before its inflow into the lake
180 (SR). The transect in the outflow comprised 7 sites, the Selenga River Mouth (SRM), the surface
181 water at 1, 3 and 5 km from the outflow (S1, S3, S5) and the bottom water at 1, 3 and 5 km from the
182 outflow (B1, B3, B5). A shoreline SPM sample from a site in Lake Baikal (Ba) not directly influenced
183 by the river, close to the point where the lake water flows into the Angara River, allowed contrasting
184 the Lake Baikal distribution exported with the Yenisei river system further downstream, before (Y1)
185 and after (Y2) the contribution of the Angara River. Furthermore, the brGDGT distribution of the
186 SPM of a mountainous river (i.e. the Irkut River; MIR) was evaluated to constrain the influence that
187 similar mountainous rivers that drain into the lake may have on the lacustrine brGDGT distribution.
188 The sites and bulk parameters for the organic matter (OM) in SR, MIR, Y1 and Y2 have been reported
189 by De Jonge et al. (2014a), while the GDGT concentrations and distributions have been reported
190 previously by De Jonge et al. (submitted).

191

192 *4.1. Environmental and bulk OM parameters*

193 The Selenga River, Lake Baikal and the Mountain River are contrasting waterbodies, as
194 reflected in both the temperature and bulk POC content, and to a lesser extent in the pH and bulk
195 $\delta^{13}\text{C}_{\text{POC}}$ values (Table 1). The inflow of Selenga River into the lake is characterized by a strong
196 gradient in water temperature (Fig. 3A), decreasing from 25 °C in the river and 22.5 °C at the river
197 mouth to 11.3 °C at S5. The bottom water is on average 6°C colder than the surface water. The coldest
198 water in the Selenga outflow system (5.2 °C) was the bottom water at 5 km distance. The Lake Baikal
199 outflow sample (Ba) had a temperature of 6.7 °C. The pH remained more constant in the outflow
200 system (Fig. 3B), with a river water pH of 8.4, while the outflow system had a pH that varied around
201 8.0 pH units. The POC varied between 6.4 and 1.1 mg/l, with the highest values in the river. A clear
202 downstream trend was observed (Fig. 3C), POC decreasing as the river flows further in Lake Baikal.
203 The POC content of the bottom water (2.5-1.1 mg/l) was always lower than for the corresponding
204 surface water, although the difference was most pronounced near the shore. The $\delta^{13}\text{C}_{\text{POC}}$ values (Fig.
205 3D) showed that the lake bottom water to have an increasingly negative value downstream (down to -
206 29.7 ‰) compared with the river SPM (-27.6 ‰). Surface water in the outflow system also had more
207 negative value of $\delta^{13}\text{C}_{\text{POC}}$ than the river, except for the sites 3 and 5km from the river mouth, that had
208 a comparable value as the river mouth (-28.6 to -28.4 ‰). The shoreline lake site had an offset, much
209 less negative $\delta^{13}\text{C}_{\text{POC}}$ value (-24.0 ‰). The mountainous Irkut River differed from the lowland Selenga
210 River, as it has the lowest POC content in the (0.1 mg/l), a much lower pH (5.6) and a lower water
211 temperature (10.8°C). On the other hand, the $\delta^{13}\text{C}_{\text{POC}}$ value of -27.8 ‰ was comparable to that in the
212 Selenga River. The Yenisei River before and after the inflow of the Angara River had a neutral pH
213 (7.5-7.9), a temperature between 11 and 12 °C, and the one measured $\delta^{13}\text{C}_{\text{POC}}$ value (-30.5 ‰) was
214 slightly more negative than for both the Selenga River and Lake Baikal systems.

215

216 *4.2. GDGT amounts and distribution*

217 The amounts and distribution of crenarchaeol and the brGDGTs are listed in Table 2. SPM of
218 the Selenga River and the river mouth had a low amount of crenarchaeol (0.3-0.5 µg/g POC; Fig. 4B),

219 which increased slightly in the outflow surface water (up to 0.5 $\mu\text{g/g}$ POC). SPM of the bottom waters
220 contained a higher amount of crenarchaeol (0.9-3 $\mu\text{g/g}$ POC), with the highest amount 3 km from the
221 delta. The lake outflow and the Yenisei River SPM had a low crenarchaeol concentration, between 0.2
222 and 0.3 $\mu\text{g/g}$ POC. The amount of brGDGTs was highest at the Selenga River Mouth (10-12 $\mu\text{g/g}$
223 POC; Fig. 4A), substantially higher than in the river itself (4 $\mu\text{g/g}$ POC), although this was measured
224 in a different month and can thus not be directly compared. The brGDGT concentration decreased in
225 the surface water of the outflow system to 0.6 $\mu\text{g/g}$ POC. The mountainous Irkut River had the highest
226 POC-normalized brGDGT concentration at 15 $\mu\text{g/g}$ POC. The brGDGTs in the bottom water varied
227 between 2 and 10 $\mu\text{g/g}$ POC, with the highest values 3 km from the Selenga River Mouth. The lake
228 outflow had the lowest amount of brGDGTs, with 0.1 $\mu\text{g/g}$ POC. The abundance in the Yenisei River
229 SPM varies between 10 and 21 $\mu\text{g/g}$ POC (De Jonge et al., 2014b).

230 Typical examples of brGDGT distributions in the SPM are shown in Fig. 5. The distribution in
231 the Selenga River Mouth (Fig. 5A) was dominated by 6-Me pentamethylated brGDGTs, and relatively
232 abundant cyclopentane-containing brGDGTs (17%). It was markedly different from that in the lake
233 outflow (Fig. 5E), which was dominated by 5-Me and 6-Me hexamethylated brGDGTs, and a lower
234 contribution of cyclopentane-containing brGDGTs (5%). The brGDGTs in the Selenga outflow
235 seemed to have intermediate distributions (Fig. 5B-D), while the mountain river had a distribution
236 comparable with that of the lake outflow (Fig. 5E-F). The distribution in the Yenisei River samples
237 (Fig. 5G-H) was dominated by 6-Me brGDGTs, the hexamethylated compounds in particular
238 dominated.

239 To study the variance in the brGDGT distribution of the lake and the rivers in its watershed
240 (Selenga and Irkut), a PCA was performed, based on the fractional abundances of brGDGTs in the
241 Selenga River, Lake Baikal and the mountainous Irkut River ($n = 10$). Most brGDGTs containing two
242 cyclopentanes, i.e. IIc, IIc', IIIc and IIIc' were excluded, as they were absent from 4 to 7 samples. The
243 first three PCs accounted for a large part of the variance (almost 90%), being 41%, 25% and 21%,
244 respectively. The first PC highlighted the correlation of a group of minor brGDGTs with positive
245 values on PC1 (Ib, Ic, IIb, IIIb, IIa', IIb', IIIb') and, to a lesser extent, brGDGT Ia (Fig. 6A). The
246 brGDGTs IIa and IIIa had negative values on PC1. PC2 revealed that brGDGT IIIa', and to a lesser

247 extent IIa' and IIb', showed a negative correlation with Ic, IIb and IIIb (Fig. 6A). BrGDGT Ia had a
248 positive score on PC3, while IIIb and IIIb' had a negative score (Fig. 6C). The similarity in the
249 brGDGT distributions at each site could be evaluated comparing their scores on the first three PCs.
250 The similar scores for the Selenga River, Selenga River Mouth and the surface water 1 km from the
251 river mouth, indicated that the brGDGT distributions were similar (Fig. 6B, D). Moving downstream
252 (S3, S5) the scores on PC1 decreased, and the scores on PC2 and, especially, PC3 increased. The
253 distribution for S3 (Fig. 5B) indeed showed increased fractional abundances of the 5-Me brGDGTs Ia,
254 IIa and IIIa, at the expense of decreased fractional abundances of 6-Me brGDGTs. This was also
255 apparent from the IR ratio (Table 2; Fig. 8A) that decreased downstream in the Selenga outflow
256 surface water. Compared with the overlying surface water, SPM of the bottom water had an increased
257 score on PC2, reflecting a decrease in brGDGT IIIa', coupled with an increase in brGDGTs IIa and
258 IIIa, that dominated the distributions of B3 and B5, respectively (Fig. 5C, 5D). This resulted in lower
259 IR values (Table 2; Fig. 8A) of the bottom water vs. the overlying surface water. The brGDGT
260 distribution in the lake outflow, was most similar to that in the mountainous Irkut River, as reflected in
261 the similar scores on PC1 and PC3. However, on PC2, the lake outflow sample had more negative
262 values, caused by a larger fractional abundance of IIIa and IIIa' (Fig. 5E).

263 The point measurement of water temperature had a significant correlation with the PCs ($p < 0.05$)
264 and were plotted a posteriori in the PCA. The water temperature values had a positive score on PC1,
265 and a negative one on PC2. The point measurements of water pH showed no significant correlation
266 with the PCs ($p < 0.05$).

267

268 **5. Discussion**

269 *5.1. Characteristics of Selenga River outflow*

270 During the sampling period (September 2011), the Selenga outflow system was characterized by
271 a warmer surface layer that extended to at least 5 km in Lake Baikal. Although a strong thermal bar is
272 usually present during the start of the summer (Sorokovikova et al., 2012), its stability generally
273 breaks down later in summer. The temperature profile shows that there was no complete stratification
274 of the Selenga River outflow water during the sampling, and partial mixing caused the bottom water

275 temperature to warm, especially near the shoreline (Fig. 3A). The $\delta^{13}\text{C}_{\text{POC}}$ values of the Selenga River
276 and River Mouth SPM (-27.5 to -28.5‰) are indicative of a dominant terrigenous signature, probably
277 derived from soil or terrigenous vegetation (typically -27‰; Michener and Lajtha, 2007). Although
278 large blooms of phytoplankton were described to be present in the mixing zone of the river inflow
279 (Sorokovikova et al.; 2012), this apparently has little impact on the $\delta^{13}\text{C}_{\text{POC}}$ values in the Selenga
280 outflow surface SPM (Fig. 3D). This may be due to the similar $\delta^{13}\text{C}$ signature of the lake
281 phytoplankton and the terrigenous organic matter (OM), as the former varies between -27 and -29.5‰
282 (Yoshii et al., 1999). Bottom water SPM, however, shows increasingly negative $\delta^{13}\text{C}_{\text{POC}}$ values,
283 possibly reflecting the slightly more negative $\delta^{13}\text{C}_{\text{POC}}$ values of lacustrine phytoplankton, compared to
284 the Selenga River, as the highest abundance of lake phytoplankton was in the bottom waters at 5 km
285 distance (Sorokovikova et al., 2012). Alternatively, this may be caused by a shift in the microbial
286 community (Maksimenko et al., 2008; Sorokovikova et al., 2012), possibly with an increased
287 contribution of organotrophic bacteria. Increased amounts of such bacteria in the bottom water have
288 been described, with maximum abundance in the bottom water at 5-10 (Maksimenko et al., 2008) or
289 14 km from the Selenga River mouth (Sorokovikova et al., 2012). They may thrive on the easily
290 hydrolysable OM entering the near-bottom layers from decomposition of the spring algae (Votintsev
291 et al., 1975). The shoreline lake SPM probably does not reflect an open water lacustrine end member.
292 Here, the POC is substantially more enriched in ^{13}C , but this probably reflects the contribution of a
293 isotopically enriched OM derived from macrophytes ($\delta^{13}\text{C}$ between -19 and -5‰; Yoshii, 1999) and
294 benthic algae ($\delta^{13}\text{C}$ between -12 and -5‰; Yoshii, 1999), as the SPM was sampled in a shallow
295 setting.

296

297 *5.2. Changing GDGT abundance in the Selenga River outflow*

298 As the Selenga River is the main source for SPM in Lake Baikal, its iGDGT and brGDGT
299 distribution is expected to contribute to the lacustrine GDGT signature. The iGDGT crenarchaeol is
300 present only in a low abundance in the river water. As it is a tracer for mesophilic Thaumarchaeota
301 (Sinninghe Damsté et al., 2002), this fits with a reported absence of archaea from the river water
302 (Maksimenko et al., 2008). The amount of archaea increases in the lake (Maksimenko et al., 2008) and

303 the increase is mimicked by a slight increase in the amount of crenarchaeol in the outflow surface
304 water (Fig. 4B). The downstream trend in the brGDGT concentration in the outflow is different, being
305 highest in the river mouth, and decreasing in the surface water of the river outflow system (Fig. 4A).
306 The decrease in brGDGT concentration in an outflow system is similar to studies where brGDGTs are
307 traced in river outflows in marine environments (e.g. Hopmans et al., 2004; Zell et al., 2014; De Jonge
308 et al., submitted). The decrease in POC-normalized concentration of riverine brGDGTs after their
309 inflow into Lake Baikal can be a result of dilution in the lacustrine environment due to in-situ
310 produced OM, or to degradation in the active biological system of the river outflow (Maksimenko et
311 al., 2008; Sorokovikova et al., 2012). However, the brGDGT concentration in the bottom water 3 and
312 5 km from the river mouth, exceeds the surface water concentration. At 3 km from the outflow, the
313 crenarchaeol concentration in the bottom water also exceeds the concentration in the surface water,
314 (i.e. a fivefold increase).

315 As no significant decrease in POC concentration was observed at B3 and B5 (Fig. 3C), we infer
316 that the increase in GDGTs at these sites is not an artifact of the normalization to POC. Although they
317 are produced by different source organisms, both crenarchaeol and brGDGTs have been described to
318 be produced in-situ in the lacustrine environment. Crenarchaeol was shown to be an abundant
319 compound in the marine environment (Sinninghe Damsté et al., 2002), but also in most medium-sized
320 and large lakes studied (Blaga et al., 2009), as it is produced by marine and freshwater
321 Thaumarchaeota (e.g. Sinninghe Damsté et al., 2002, Pearson et al., 2004). These NH₃ oxidizing
322 Archaea have a preferred niche in the oxycline/thermocline and nitrocline of the water column of lakes
323 (e.g. Pouliot et al., 2009; Lirós et al., 2010; Auguet et al., 2011; Blaga et al., 2011; Schouten et al.,
324 2012; Woltering et al., 2012; Buckles et al., 2013), where they can outcompete NH₃-oxidizing
325 bacteria. The presence of an increased supply of decomposed algal matter in near-bottom water layers,
326 resulting in increased NH₃ concentration (e.g. Jewell and McCarty, 1971), can possibly fuel the
327 crenarchaeol production in these deeper layers. Also in-situ production of brGDGTs has been
328 observed in lacustrine environments, based on shifts in the distribution and increased abundance (e.g.
329 Tierney and Russell, 2009). BrGDGT producers in soils are believed to be heterotrophic bacteria (e.g.
330 Pancost and Sinninghe Damsté, 2003; Weijers et al., 2010). Thus, the in-situ production at these sites

331 can possibly be linked to enhanced heterotrophic activity in the bottom water. Indeed, the amount of
332 organotrophic bacteria was shown to be slightly increased in the bottom water at 5-10 km from the
333 Selenga River Mouth (Maksimenko et al., 2008).

334 To trace the transport of bacterial OM (soil-derived and/or river-derived) in the marine system,
335 the BIT index was developed (Hopmans et al., 2004), expressed as the abundance of the non-
336 cyclopentane containing brGDGTs, relative to the abundance of crenarchaeol. In the Selenga outflow,
337 the concentration of brGDGTs is the dominant driver for the changing BIT index (Fig. 4C), as it
338 correlates with the amount of brGDGTs ($r^2 = 0.63$). The index decreases with increasing distance from
339 the river mouth. The decrease is to be expected, and has been observed in front of several river
340 outflows in the marine system (e.g. Hopmans et al., 2004; Zell et al., 2014; De Jonge et al., submitted).
341 The slightly increased values in the bottom water reflect the increased amount of brGDGTs there
342 compared with the surface outflow water. The value for the lake outflow (0.22), is lower than reported
343 values for sediment cores from the lake, where Holocene values vary between 0.3 and 0.6 (Fietz et al.,
344 2011), although the difference can still be related to interlaboratory differences that can influence the
345 values obtained (Schouten et al., 2009, 2013). Overall, the brGDGT abundance decreases significantly
346 in the river outflow, while the crenarchaeol abundance remains stable. However, in-situ production of
347 both crenarchaeol and brGDGTs within the bottom water is likely. Based on the distance from the
348 river mouth, the increase is possibly fueled by increased degradation of phytoplankton material in the
349 bottom water (Maksimenko et al., 2008; Sorokovikova et al., 2012).

350

351 *5.3. Changing brGDGT distributions in the outflow of Selenga River and Lake Baikal*

352 The distribution in the Selenga River and the River Mouth is distinctly different from the
353 shoreline lake SPM that is presumably transported further downstream to the Yenisei River (Fig. 5A,
354 F). Downstream changes in the fractional abundance of all 15 brGDGTs can be seen in the
355 distributions (e.g. Fig. 5A-E) and the PCA based on the fractional abundances (Fig. 6). To evaluate
356 whether the downstream changes are due exclusively to linear mixing between a riverine (SRM) and
357 lacustrine (Ba) endmember, we plotted the data in a triplot based on the fractional abundance of three
358 major brGDGTs (Fig. 7). It captures the majority of the variance, as the three brGDGTs have high

359 scores on the first three PCs (brGDGTs IIIa, IIIa' and Ia, respectively). In the case that linear mixing is
360 the only mechanism causing the shift in brGDGT distribution in the Yenisei outflow, and if the two
361 endmembers (lacustrine, riverine) are well represented by the Ba and SRM samples, distributions that
362 result from conservative mixing of these endmembers alone should fall on the thick black line in Fig.
363 7. However, both surface and bottom water SPM along the Selenga outflow transect do not fall on this
364 mixing line. Compared with the Selenga River Mouth, the surface outflow water shows a stepwise
365 increase in the fractional abundance of IIIa, and a decrease for Ia and IIIa'. Bottom water SPM also
366 shows an increase in IIIa (in B1, but especially B3 and B5) and a decrease in Ia and IIIa'. The increase
367 in IIIa correlates well with a decrease in Ib, IIa' and IIb' ($r^2 = 0.57, 0.90$ and 0.96 , respectively).
368 Possible explanations for the offset from the linear mixing line are preferential degradation of a
369 brGDGT pool that was enriched in the Ia, Ib, IIa', IIb' and IIIa', or the in-situ production of IIIa in the
370 outflow. Although the brGDGT distribution in the river reflects a mixture dominated by in-situ
371 produced brGDGTs, a small contribution from soil-derived brGDGTs could be present. As the riverine
372 OM will be more easily degraded (e.g. Blair and Aller, 2012), a soil-derived brGDGT sub-pool could
373 become more dominant during degradation of the OM delivered by the river. However, especially in
374 the bottom water, in-situ production of brGDGTs is probable, based on the increased brGDGT
375 concentration. In-situ production of brGDGTs in lacustrine systems has been described in several
376 studies, both for the lake water column or sediments. Unfortunately, these studies do not allow
377 discussing changes in the abundance of 5- and 6-Me brGDGTs separately, although the relative
378 abundance of the 6-Me brGDGTs, expressed as the isomer ratio (IR, Eq. 1; Fig. 8A) shows strong
379 changes downstream. The contrasting behavior of the 5- and 6-Me brGDGTs is also reflected in their
380 dissimilar scores in the PCA (Fig. 6A, C) and ideally they should therefore be studied separately when
381 discussing aquatic brGDGT distributions.

382 Loomis et al. (2014) have shown that cyclopentane-containing brGDGTs and penta- and
383 hexamethylated brGDGTs can be produced in the water column of a temperate lake, both in the
384 surface water and bottom water. A study (Buckles et al., 2014a) of the partially anoxic Lake Challa
385 showed that penta- and hexamethylated brGDGTs were produced in the oxic part of the water column,
386 but mainly in the suboxic part, and found some evidence suggesting their production in sediments.

387 Furthermore, the authors showed that in-situ production of IIIa (and/or IIIa') and IIb (and/or IIb')
388 occurred in the Loch Lomond (Scotland), although they could not pinpoint whether the production
389 was localized in the sediments or in the water column. Indications were also found for in-situ
390 production of penta- and hexamethylated brGDGTs in the suboxic sediments of Sand Pond (USA;
391 Tierney et al., 2012). Overall, previous studies agree on the production of highly methylated and
392 cyclopentane-containing brGDGTs by brGDGT producing bacteria occupying a variety of niches
393 within the lacustrine environment. This is in line with the postulated in-situ production of IIIa in the
394 Selenga Outflow. As IIc, IIc', IIIc and IIIc' were below detection limit in 4-7 samples, an increase in
395 their fractional abundance could not be established.

396 We have shown that the distribution in the Selenga River outflow cannot be explained solely by
397 mixing of a riverine end member and a lacustrine endmember. However, the source of the brGDGT
398 distribution in the lacustrine endmember (Ba) still remains to be constrained. Besides being in-situ
399 produced, a second possible source for the brGDGT distribution in Lake Baikal is brGDGTs delivered
400 by other rivers that account up to 50% of its water volume, and up to 20% of the total suspended solids
401 (Votintsev et al., 1985). Although we did not sample the SPM from such rivers within the watershed,
402 we sampled the brGDGTs in the headwater of Irkut River, a mountainous river in a nearby and similar
403 location (Fig. 2), for which we postulate that it should approach the brGDGT distribution in the
404 northern and eastern rivers of the watershed of the lake and in the headwaters of Selenga River. The
405 distribution in the SPM of this mountainous river and that of the lacustrine endmember are rather
406 similar (Fig. 5E, F), as confirmed by their similar scores on PC1 and PC3 (Fig. 6B, D). Although
407 linear mixing (Fig. 7, dotted line) between the distributions in the mountain river and the Selenga
408 River, approaches the lacustrine distribution, a small decrease in the fractional abundance of Ia and
409 IIIa, or an increase in IIa', IIb' and IIIa' is needed to explain the distribution in the lake.
410 Unfortunately, it is impossible to determine whether preferential degradation or in-situ production of
411 brGDGTs is the more probable explanation.

412 Another issue is how the lake water affects the brGDGT composition of downstream Yenisei
413 river water. The distribution in the latter (Figs. 5G-H; Fig. 7) is distinctly different from that of the
414 lake outflow (Fig. 5E; Fig. 7) with a much larger fractional abundance of 6-Me brGDGTs. This

415 dominance becomes even more evident after the inflow of the Angara River, that drains lake water
416 into the Yenisei River. The signature of the latter (Figs. 5G-H; Fig. 7) is more similar to that of the
417 Selenga River (Fig. 5A; Fig. 7). This indicates that riverine in-situ production of brGDGTs, a process
418 reported to be dominant for brGDGT distribution in the whole Yenisei River (De Jonge et al., 2014b),
419 already operates in the Angara River. This is confirmed by the two orders of magnitude higher
420 concentration of brGDGTs in the Yenisei River SPM than in the lake. As the lacustrine brGDGT
421 distribution is not present in downstream SPM, we postulate that brGDGTs sourced from the
422 watershed of the lake, including the Selenga River watershed, should not contribute significantly to
423 the brGDGT distribution transported by the Yenisei River and delivered to the marine system.

424

425 *5.4. Environmental controls on brGDGT distributions in the Selenga River outflow*

426 The lacustrine sedimentary brGDGT distribution of lakes covering a latitudinal or altitudinal
427 range responds to overlying air temperature (e.g. Tierney et al., 2010; Sun et al., 2011; Pearson et al.,
428 2011; Loomis et al., 2012) and to seasonal changes in water temperature (e.g. Loomis et al., 2014a).
429 Similar to the soil calibrations, these authors found an increase in Ia in warmer lake settings, and a
430 decrease of penta- and hexamethylated brGDGTs. Furthermore, some authors found evidence that pH
431 has an influence on the brGDGT distributions (Tierney et al., 2010; Loomis et al., 2014a; Schoon et
432 al., 2013), while others found no indication of this (Pearson et al., 2011; Loomis et al., 2014b).

433 The measured environmental parameters (point measurements of pH and water temperature)
434 have been plotted a posteriori in the ordination space (Fig. 6A), in the case of significant correlation (p
435 < 0.05) with the PC. The value of this vector in the ordination space is based on the correlations
436 between the values of the environmental parameters at the sites, with the site scores on the PCs and
437 thus revealing their relationship with individual brGDGTs. Their direction indicates for which samples
438 their values are increased, and with which brGDGT correlations are present. The measured water
439 temperature values correlated significantly with the site scores on the first two PCs ($p < 0.02$). The
440 increase in fractional abundance of IIa and IIIa in the Selenga outflow results in a negative correlation
441 with the measured water temperature as the temperature decreases strongly in the Selenga River
442 outflow ($r^2 = 0.34$ and 0.64 , respectively). The measured water temperature follows a positive

443 correlation with the fractional abundance of Iia' and Iib' ($r^2 = 0.75$ and 0.59 , respectively). The
444 measured pH values do not correlate significantly with the main trends in the brGDGT distribution.

445 As the change in brGDGT distribution seems to be related to water temperature, we evaluate the
446 performance of a MST calibration based on the fractional abundance of lacustrine brGDGTs (Pearson
447 et al., 2011; Eq. 6). However, to date, no lake calibration has been based on the extended dataset of 15
448 brGDGTs. Although this calibration excluded IIIa' (E. Pearson, personal communication), it included
449 the pentamethylated 6-Me brGDGTs, that vary strongly in the Selenga River outflow. The
450 reconstructed MST for the outflow surface water varies between 18 and 20 °C (Fig. 8B). The bottom
451 water values, except for sample B3 (17-19 °C; Fig. 8B), reflect the same temperature as the surface
452 values, within the RSME of the calibration (2 °C; Pearson et al., 2011). As the absolute water
453 temperature decreases from 25 to 12 °C (Fig. 3A), the reconstructed temperature values only give a
454 muted response to this large decrease in temperature. The muted response to the temperature shift
455 measured in-situ is a phenomenon also recognized by Loomis et al. (2014a). Although major shifts in
456 the fractional abundance of the penta- and hexamethylated brGDGTs occur in the Selenga outflow,
457 this is thus not reflected in the reconstructed temperature. This is caused by the combination of 5- and
458 6-Me pentamethylated brGDGTs. The Pearson et al. (2011) calibration is based on the air MST, that
459 varies between 10.6 and 15.4 °C in the Lake Baikal watershed, based on 5 climate stations (De Jonge
460 et al., 2014a). In this respect, the reconstructed MST overestimates the MST measured in the
461 watershed.

462 Although the brGDGT distribution at B3 is generally comparable with that in the other river
463 outflow samples, it results in a substantially lower reconstructed temperature (15 °C; decrease of 4
464 °C), because of the larger fractional abundance of Iia. This decrease in reconstructed temperature is
465 not related to a decrease in water temperature (Fig. 3A). As strong shifts in the abundance of nutrients
466 have shown to cause shifts in brGDGT distributions that are probably related to shifts in the microbial
467 community (Loomis et al., 2014a), the low reconstructed temperature values can be due to the
468 presence of a distinct brGDGT producing bacterial community associated with an increased amount of
469 phytoplankton-derived detritus.

470 We also tested the performance of the most recent soil pH calibration (De Jonge et al., 2014b) in
471 this aquatic setting. To this end, a pH reconstruction was performed using the CBT', a ratio based on a
472 dataset where all 15 brGDGTs were quantified separately. The reconstructed pH (Eqs. 3 and 4) has
473 absolute values (7-7.5) that only slightly underestimate the pH measured in the river and lake water
474 (7.9-8.9), especially when taking the residual error of 0.5 pH units into account. The reconstructed pH
475 shows a slight decrease in the Selenga outflow surface water, that is an intensified response compared
476 with the measured values. In the outflow system, sample B3, with a reconstructed pH of 6.5 has the
477 largest offset with the measured pH (7.9). The brGDGT signature at the site, which is most probably
478 influenced by an in-situ produced brGDGTs, thus fails to reconstruct the in-situ measured pH.
479 Furthermore, although the mountainous Irkut River has a lower pH than the lake, the brGDGT
480 distributions are fairly similar. Overall, this environmental dataset indicates a limited influence of pH
481 on the aquatic brGDGT distributions, as all reconstructed pH values vary around 7-7.5, and do not
482 correlate with the measured values.

483

484 **6. Conclusions**

485 We have traced the Selenga River brGDGT signature during its outflow into Lake Baikal and
486 compared the brGDGT signature with the outflow from Lake Baikal and the brGDGT distribution in
487 the Yenisei River SPM. The signature for the Selenga River is characterized by a dominance of 6-Me
488 brGDGTs IIa' and IIIa' (Selenga River endmember). Although the Selenga River delivers the largest
489 part of the suspended matter to Lake Baikal, the brGDGT distribution delivered by it is drastically
490 different from the lacustrine brGDGT distribution, that shows an increased amount of brGDGT IIa and
491 IIIa (lacustrine endmember). As linear mixing of these two endmembers can not explain the shifts in
492 the Selenga Outflow brGDGT distribution, we postulate that in-situ production of brGDGTs or
493 preferential degradation takes place in the Selenga outflow. The presence of active archaeal/bacterial
494 communities in dynamic river outflow systems in lakes can thus possibly influence the brGDGT
495 distribution encountered in lake sediments, especially as the extent of these outflow systems changes
496 within and in between years. The brGDGT distribution of the lacustrine endmember can be produced
497 in-situ in the lake, or can be caused by a contribution from brGDGTs derived from mountainous rivers

498 in the Selenga watershed, that is dominated by non-cyclopentane containing penta- and
499 hexamethylated brGDGTs (mountain river endmember).

500 Furthermore, the brGDGT distribution in the Yenisei River, both before and after the inflow of
501 the Angara River that drains Lake Baikal, was studied. This distribution was found to be dissimilar to
502 the lacustrine endmember, indicating only a limited contribution from the lacustrine brGDGTs to the
503 Yenisei River distribution. The dominance of 6-Me brGDGTs, similar to the brGDGT distribution
504 encountered in the Selenga River, indicated that the lacustrine brGDGT signature is overwritten by in-
505 situ produced riverine brGDGTs (De Jonge et al., 2014a; submitted). The watershed of Lake Baikal
506 thus does not contribute significantly to the brGDGT pool transported by the Yenisei River. Large
507 lakes are a common feature in river watersheds globally and this study indicates that their presence can
508 result in the absence of brGDGTs sourced in areas upstream of these lakes. This has to be taken into
509 account when interpreting palaeoclimate reconstructions based on brGDGTs from river fan sediments.
510 This study therefore has several implications for the use of lipid-based palaeoclimate proxies, both in
511 lacustrine and marine sediments.

512

513

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523 **References**

- 524 Auguet, J.-C., Nomokonova, N., Camarero, L., Casamayor, E.O., 2011. Seasonal changes of freshwater
525 ammonia-oxidizing Archaeal assemblages and nitrogen species in oligotrophic alpine lakes. *Applied*
526 *Environmental Microbiology* 77, 1937–1945.
- 527 Bendle, J.A., Weijers, J.W.H., Maslin, M.A., Sinninghe Damsté, J.S., Schouten, S., Hopmans, E.C., Boot, C.S.,
528 Pancost, R.D., 2010. Major changes in glacial and Holocene terrestrial temperatures and sources of organic
529 carbon recorded in the Amazon fan by tetraether lipids. *Geochemistry Geophysics Geosystems* 11,
530 Q12007.
- 531 Blaga, C.I., Reichart, G.J., Heiri, O., Sinninghe Damsté, J.S., 2009. Tetraether membrane lipid distributions in
532 water-column particulate matter and sediments: a study of 47 European lakes along a north–south transect.
533 *Journal of Paleolimnology* 41, 523–540.
- 534 Blaga, C.I., Reichart, G.-J., Vissers, E.W., Lotter, A.F., Anselmetti, F.S., Sinninghe Damsté, J.S., 2011. Seasonal
535 changes in glycerol dialkyl glycerol tetraether concentrations and fluxes in a perialpine lake: Implications
536 for the use of the TEX₈₆ and BIT proxies. *Geochimica et Cosmochimica Acta* 75, 6416–6428.
- 537 Blair, N.E., Aller, R.C., 2012. The fate of terrestrial organic carbon in the marine environment. *Annual Review*
538 *of Marine Science* 4, 401–423.
- 539 Blyth, A.J., Schouten, S., 2013. Calibrating the glycerol dialkyl glycerol tetraether temperature signal in
540 speleothems. *Geochimica et Cosmochimica Acta* 109, 312–328.
- 541 Buckles, L.K., Villanueva, L., Weijers, J.W.H., Verschuren, D., Sinninghe Damsté, J.S., 2013. Linking
542 isoprenoidal GDGT membrane lipid distributions with gene abundances of ammonia-oxidizing
543 Thaumarchaeota and uncultured crenarchaeotal groups in the water column of a tropical lake (Lake Challa,
544 East Africa). *Environmental Microbiology* 15, 2445–2462.
- 545 Buckles, L.K., Weijers, J.W.H., Verschuren, D., Sinninghe Damsté, J.S., 2014a. Sources of core and intact
546 branched tetraether membrane lipids in the lacustrine environment: Anatomy of Lake Challa and its
547 catchment, equatorial East Africa. *Geochimica et Cosmochimica Acta* 140, 106–126.
- 548 Buckles, L.K., Weijers, J.W.H., Tran, X.-M., Waldron, S., Sinninghe Damsté, J.S., 2014. Provenance of
549 tetraether membrane lipids in a large temperate lake (Loch Lomond, UK): implications for GDGT-based
550 palaeothermometry. *Biogeosciences* 11, 5539–5563.
- 551 De Jonge, C., Hopmans, E.C., Stadnitskaia, A., Rijpstra, W.I.C., Hofland, R., Tegelaar, E., Sinninghe Damsté,
552 J.S., 2013. Identification of novel penta- and hexamethylated branched glycerol dialkyl glycerol tetraethers
553 in peat using HPLC–MS², GC–MS and GC–SMB-MS. *Organic Geochemistry* 54, 78–82.
- 554 De Jonge, C., Stadnitskaia, A., Hopmans, E.C., Cherkashov, G., Fedotov, A., Sinninghe Damsté, J.S., 2014a. In
555 situ produced branched glycerol dialkyl glycerol tetraethers in suspended particulate matter from the
556 Yenisei River, Eastern Siberia. *Geochimica et Cosmochimica Acta* 125, 476–491.
- 557 De Jonge, C., Hopmans, E.C., Zell, C.I., Kim, J.-H., Schouten, S., Sinninghe Damsté, J.S., 2014b. Occurrence
558 and abundance of 6-methyl branched glycerol dialkyl glycerol tetraethers in soils: Implications for
559 palaeoclimate reconstruction. *Geochimica et Cosmochimica Acta* 141, 97–112.
- 560 De Jonge, C., Stadnitskaia, A., Hopmans, E.C., Cherkashov, G., Fedotov, A., Streletskaia, I.D., Vasiliev, A.,
561 Sinninghe Damsté, J.S. Drastic changes in the distribution of branched tetraether lipids in suspended matter
562 and sediments from the Yenisei River and Kara Sea (Siberia): Implications for the use of brGDGT-based
563 proxies in coastal marine sediments. Submitted to *Geochimica et Cosmochimica Acta*.
- 564 Fietz, S., Martínez-García, A., Huguet, C., Rueda, G., Rosell-Melé, A., 2011. Constraints in the application of
565 the Branched and Isoprenoid Tetraether index as a terrestrial input proxy. *Journal of Geophysical Research:*
566 *Oceans* 116, C10032.
- 567 Hopmans, E.C., Weijers, J.W.H., Schefuß, E., Herfort, L., Sinninghe Damsté, J.S., Schouten, S., 2004. A novel
568 proxy for terrestrial organic matter in sediments based on branched and isoprenoid tetraether lipids. *Earth*
569 *and Planetary Science Letters* 224, 107–116.
- 570 Huguet, C., Hopmans, E.C., Febo-Ayala, W., Thompson, D.H., Sinninghe Damsté, J.S., Schouten, S., 2006. An
571 improved method to determine the absolute abundance of glycerol dibiphytanyl glycerol tetraether lipids.
572 *Organic Geochemistry* 37, 1036–1041.
- 573 Jewell, W.J., McCarty, P.L., 1971. Aerobic decomposition of algae. *Environmental Science and Technology* 5,
574 1023–1031.

- 575 Kim, J.-H., Zell, C., Moreira-Turcq, P., Perez, M.A.P., Abril, G., Mortillaro, J.-M., Weijers, J.W.H., Meziane,
576 T., Sinninghe Damsté, J.S., 2012. Tracing soil organic carbon in the lower Amazon River and its tributaries
577 using GDGT distributions and bulk organic matter properties. *Geochimica et Cosmochimica Acta* 90, 163–
578 180.
- 579 Lliros, M., Gich, F., Plasencia, A., Auguet, J.-C., Darchambeau, F., Casamayor, E.O., Descy, J.-P., Borrego, C.,
580 2010. Vertical distribution of ammonia-oxidizing Crenarchaeota and methanogens in the epipelagic waters
581 of Lake Kivu (Rwanda-Democratic Republic of the Congo). *Applied and Environmental Microbiology* 76,
582 6853–6863.
- 583 Loomis, S.E., Russell, J.M., Sinninghe Damsté, J.S., 2011. Distributions of branched GDGTs in soils and lake
584 sediments from western Uganda: Implications for a lacustrine paleothermometer. *Organic Geochemistry*
585 42, 739–751.
- 586 Loomis, S.E., Russell, J.M., Ladd, B., Street-Perrott, F.A., Sinninghe Damsté, J.S., 2012. Calibration and
587 application of the branched GDGT temperature proxy on East African lake sediments. *Earth and Planetary*
588 *Science Letters* 357-358, 277–288.
- 589 Loomis, S.E., Russell, J.M., Eggermont, H., Verschuren, D., Sinninghe Damsté, J.S., 2014a. Effects of
590 temperature, pH and nutrient concentration on branched GDGT distributions in East African lakes:
591 Implications for paleoenvironmental reconstruction. *Organic Geochemistry* 66, 25–37.
- 592 Loomis, S.E., Russell, J.M., Heures, A.M., D’Andrea, W.J., Sinninghe Damsté, J.S., 2014b. Seasonal
593 variability of branched glycerol dialkyl glycerol tetraethers (brGDGTs) in a temperate lake system.
594 *Geochimica et Cosmochimica Acta* 144, 173-187.
- 595 Maksimenko, S.Y., Zenskaya, T.I., Pavlova, O.N., Ivanov, V.G., Buryukhaev, S.P., 2008. Microbial community
596 of the water column of the Selenga River-Lake Baikal biogeochemical barrier. *Microbiology* 77, 587–594.
- 597 Michener R. H. and Lajtha K. (Eds) 2007. *Stable isotopes in ecology and environmental science.*, Blackwell
598 Scientific Publications, Oxford, UK.
- 599 Niemann, H., Stadnitskaia, A., Wirth, S.B., Gilli, A., Anselmetti, F.S., Sinninghe Damsté, J.S., Schouten, S.,
600 Hopmans, E.C., Lehmann, M.F., 2012. Bacterial GDGTs in Holocene sediments and catchment soils of a
601 high Alpine lake: application of the MBT/CBT-paleothermometer. *Climate of the Past* 8, 889–906.
- 602 Pancost, R.D., Sinninghe Damsté, J.S., 2003. Carbon isotopic compositions of prokaryotic lipids as tracers of
603 carbon cycling in diverse settings. *Chemical Geology* 195, 29–58.
- 604 Pearson, A., Huang, Z., Ingalls, A.E., Romanek, C.S., Wiegel, J., Freeman, K.H., Smittenberg, R.H., Zhang,
605 C.L., 2004. Nonmarine Crenarchaeol in Nevada Hot Springs. *Applied Environmental Microbiology* 70,
606 5229–5237.
- 607 Pearson, E.J., Juggins, S., Talbot, H.M., Weckstrom, J., Rosen, P., Ryves, D.B., Roberts, S.J., Schmidt, R., 2011.
608 A lacustrine GDGT-temperature calibration from the Scandinavian Arctic to Antarctic: Renewed potential
609 for the application of GDGT-paleothermometry in lakes. *Geochimica et Cosmochimica Acta* 75, 6225–
610 6238.
- 611 Peterse, F., Nicol, G.W., Schouten, S., Sinninghe Damsté, J.S., 2010. Influence of soil pH on the abundance and
612 distribution of core and intact polar lipid-derived branched GDGTs in soil. *Organic Geochemistry* 41,
613 1171–1175.
- 614 Peterse, F., Prins, M.A., Beets, C.J., Troelstra, S.R., Zheng, H., Gu, Z., Schouten, S., Sinninghe Damsté, J.S.,
615 2011. Decoupled warming and monsoon precipitation in East Asia over the last deglaciation. *Earth and*
616 *Planetary Science Letters* 301, 256–264.
- 617 Pitcher, A., Hopmans, E.C., Schouten, S., Sinninghe Damsté, J.S., 2009. Separation of core and intact polar
618 archaeal tetraether lipids using silica columns: Insights into living and fossil biomass contributions.
619 *Organic Geochemistry* 40, 12–19.
- 620 Pouliot, J., Galand, P.E., Lovejoy, C., Vincent, W.F., 2009. Vertical structure of archaeal communities and the
621 distribution of ammonia monooxygenase A gene variants in two meromictic High Arctic lakes.
622 *Environmental Microbiology* 11, 687–699.
- 623 Schoon, P.L., de Kluijver, A., Middelburg, J.J., Downing, J.A., Sinninghe Damsté, J.S., Schouten, S., 2013.
624 Influence of lake water pH and alkalinity on the distribution of core and intact polar branched glycerol
625 dialkyl glycerol tetraethers (GDGTs) in lakes. *Organic Geochemistry* 60, 72–82.

- 626 Schouten, S., Hopmans, E.C., Pancost, R.D., Sinninghe Damsté, J.S., 2000. Widespread occurrence of
627 structurally diverse tetraether membrane lipids: Evidence for the ubiquitous presence of low-temperature
628 relatives of hyperthermophiles. *Proceedings of the National Academy of Sciences USA* 97, 14421–14426.
- 629 Schouten, S., Hugué, C., Hopmans, E.C., Kienhuis, M.V.M., Sinninghe Damsté, J.S., 2007. Analytical
630 methodology for TEX₈₆ paleothermometry by high-performance liquid chromatography/atmospheric
631 pressure chemical ionization-mass spectrometry. *Analytical Chemistry* 79, 2940–2944.
- 632 Schouten, S., Hopmans, E.C., van der Meer, J., Mets, A., Bard, E., Bianchi, T.S., Diefendorf, A., Escala, M.,
633 Freeman, K.H., Furukawa, Y., Hugué, C., Ingalls, A., Menot-Combes, G., Nederbragt, A.J., Oba, M.,
634 Pearson, A., Pearson, E.J., Rosell-Mele, A., Schaeffer, P., Shah, S.R., Shanahan, T.M., Smith, R.W.,
635 Smittenberg, R., Talbot, H.M., Uchida, M., Van Mooy, B.A.S., Yamamoto, M., Zhang, Z.H., Sinninghe
636 Damsté, J.S., 2009. An interlaboratory study of TEX₈₆ and BIT analysis using high-performance liquid
637 chromatography-mass spectrometry. *Geochemistry Geophysics Geosystems* 10, 13.
- 638 Schouten, S., Rijpstra, W.I.C., Durisch-Kaiser, E., Schubert, C.J., Sinninghe Damsté, J.S., 2012. Distribution of
639 glycerol dialkyl glycerol tetraether lipids in the water column of Lake Tanganyika. *Organic Geochemistry*
640 53, 34–37.
- 641 Schouten, S., Hopmans, E.C., Rosell-Melé, A., Pearson, A., Adam, P., Bauersachs, T., Bard, E., Bernasconi,
642 S.M., Bianchi, T.S., Brocks, J.J., Carlson, L.T., Castañeda, I.S., Derenne, S., Selver, A.D., Dutta, K.,
643 Eglinton, T., Fosse, C., Galy, V., Grice, K., Hinrichs, K.-U., Huang, Y., Hugué, A., Hugué, C., Hurley, S.,
644 Ingalls, A., Jia, G., Keely, B., Knappy, C., Kondo, M., Krishnan, S., Lincoln, S., Lipp, J., Mangelsdorf, K.,
645 Martínez-García, A., Ménot, G., Mets, A., Mollenhauer, G., Ohkouchi, N., Ossebaar, J., Pagani, M.,
646 Pancost, R.D., Pearson, E.J., Peterse, F., Reichart, G.-J., Schaeffer, P., Schmitt, G., Schwark, L., Shah,
647 S.R., Smith, R.W., Smittenberg, R.H., Summons, R.E., Takano, Y., Talbot, H.M., Taylor, K.W.R., Tarozo,
648 R., Uchida, M., van Dongen, B.E., Van Mooy, B.A.S., Wang, J., Warren, C., Weijers, J.W.H., Werne, J.P.,
649 Woltering, M., Xie, S., Yamamoto, M., Yang, H., Zhang, C.L., Zhang, Y., Zhao, M., Sinninghe Damsté,
650 J.S., 2013. An interlaboratory study of TEX₈₆ and BIT analysis of sediments, extracts, and standard
651 mixtures. *Geochemistry Geophysics Geosystems* 14, 5263–5285.
- 652 Sinninghe Damsté, J.S., Hopmans, E.C., Pancost, R.D., Schouten, S., Geenevasen, J.A.J., 2000. Newly
653 discovered non-isoprenoid glycerol dialkyl glycerol tetraether lipids in sediments. *Chemical*
654 *Communications* 1683–1684.
- 655 Sinninghe Damsté, J.S., Schouten, S., Hopmans, E.C., van Duin, A.C.T., Geenevasen, J.A.J., 2002.
656 Crenarchaeol: the characteristic core glycerol dibiphytanyl glycerol tetraether membrane lipid of
657 cosmopolitan pelagic crenarchaeota. *Journal of Lipid Research* 43, 1641–1651.
- 658 Sinninghe Damsté, J.S., Ossebaar, J., Abbas, B., Schouten, S., Verschuren, D., 2009. Fluxes and distribution of
659 tetraether lipids in an equatorial African lake: Constraints on the application of the TEX₈₆
660 palaeothermometer and BIT index in lacustrine settings. *Geochimica et Cosmochimica Acta* 73, 4232–
661 4249.
- 662 Sinninghe Damsté, J.S., Rijpstra, W.I.C., Hopmans, E.C., Weijers, J.W.H., Foesel, B.U., Overmann, J., Dedysh,
663 S.N., 2011. 13, 16-dimethyloctacosanedioic acid (iso-diabolic acid), a common membrane-spanning lipid
664 of Acidobacteria subdivisions 1 and 3. *Applied and Environmental Microbiology* 77, 4147–4154.
- 665 Sinninghe Damsté, J.S., Rijpstra, W.I.C., Hopmans, E.C., Foesel, B.U., Wüst, P.K., Overmann, J., Tank, M.,
666 Bryant, D.A., Dunfield, P.F., Houghton, K., Stott, M.B., 2014. Ether- and ester-bound iso-diabolic acid and
667 other lipids in members of Acidobacteria subdivision 4. *Applied and Environmental Microbiology* 80,
668 5207–5218.
- 669 Sorokovikova, L.M., Popovskaya, G.I., Belykh, O.I., Tomberg, I.V., Maksimenko, S.Y., Bashenkhaeva, N.V.,
670 Ivanov, V.G., Zemskaya, T.I., 2012. Plankton composition and water chemistry in the mixing zone of the
671 Selenga River with Lake Baikal. *Hydrobiologia* 695, 329–341.
- 672 Sun, Q., Chu, G.Q., Liu, M.M., Xie, M.M., Li, S.Q., Ling, Y.A., Wang, X.H., Shi, L.M., Jia, G.D., Lu, H.Y.,
673 2011. Distributions and temperature dependence of branched glycerol dialkyl glycerol tetraethers in recent
674 lacustrine sediments from China and Nepal. *Journal of Geophysical Research-Biogeosciences* 116, 1–12.
- 675 Tierney, J.E., Russell, J.M., 2009. Distributions of branched GDGTs in a tropical lake system: Implications for
676 lacustrine application of the MBT/CBT paleoproxy. *Organic Geochemistry* 40, 1032–1036.
- 677 Tierney, J.E., Russell, J.M., Eggermont, H., Hopmans, E.C., Verschuren, D., Sinninghe Damsté, J.S., 2010.
678 Environmental controls on branched tetraether lipid distributions in tropical East African lake sediments.
679 *Geochimica et Cosmochimica Acta* 74, 4902–4918.

680 Tierney, J.E., Schouten, S., Pitcher, A., Hopmans, E.C., Sinninghe Damsté, J.S., 2012. Core and intact polar
681 glycerol dialkyl glycerol tetraethers (GDGTs) in Sand Pond, Warwick, Rhode Island (USA): Insights into
682 the origin of lacustrine GDGTs. *Geochimica et Cosmochimica Acta* 77, 561–581.

683 Votintsev, K.K., 1985. Main features of the hydrochemistry of Lake Baikal. *Water Resources* 12, 106–116.

684 Votintsev, K.K., Mescheryakova, A.I., Popovskaya, G.I., 1975. Cycle of Organic Matter in Lake Baikal. Nauka,
685 Novosibirsk.

686 Weijers, J.W.H., Schouten, S., Hopmans, E.C., Geenevasen, J.A.J., David, O.R.P., Coleman, J.M., Pancost,
687 R.D., Sinninghe Damsté, J.S., 2006. Membrane lipids of mesophilic anaerobic bacteria thriving in peats
688 have typical archaeal traits. *Environmental Microbiology* 8, 648–657.

689 Weijers, J.W.H., Schouten, S., van den Donker, J.C., Hopmans, E.C., Sinninghe Damsté, J.S., 2007.
690 Environmental controls on bacterial tetraether membrane lipid distribution in soils. *Geochimica et*
691 *Cosmochimica Acta* 71, 703–713.

692 Weijers, J.W.H., Schefuss, E., Schouten, S., Sinninghe Damsté, J.S., 2007. Coupled thermal and hydrological
693 evolution of tropical Africa over the last deglaciation. *Science* 315, 1701–1704.

694 Weijers, J.W.H., Panoto, E., van Bleijswijk, J., Schouten, S., Rijpstra, W.I.C., Balk, M., Stams, A.J.M.,
695 Sinninghe Damsté, J.S., 2009. Constraints on the biological source(s) of the orphan branched tetraether
696 membrane lipids. *Geomicrobiology Journal* 26, 402–414.

697 Weijers, J.W.H., Wiesenberg, G.L.B., Bol, R., Hopmans, E.C., Pancost, R.D., 2010. Carbon isotopic
698 composition of branched tetraether membrane lipids in soils suggest a rapid turnover and a heterotrophic
699 life style of their source organism(s). *Biogeosciences* 7, 2959–2973.

700 Woltering, M., Werne, J.P., Kish, J.L., Hicks, R., Sinninghe Damsté, J.S., Schouten, S., 2012. Vertical and
701 temporal variability in concentration and distribution of thaumarchaeotal tetraether lipids in Lake Superior
702 and the implications for the application of the TEX₈₆ temperature proxy. *Geochimica et Cosmochimica Acta*
703 87, 136–153.

704 Yoshii, K., 1999. Stable isotope analyses of benthic organisms in Lake Baikal. *Hydrobiologia* 411, 145–159.

705 Yoshii, K., Melnik, N.G., Timoshkin, O.A., Bondarenko, N.A., Anoshko, P.N., Yoshioka, T., Wada, E., 1999.
706 Stable isotope analyses of the pelagic food web in Lake Baikal. *Limnology and Oceanography* 44, 502–
707 511.

708 Zell, C., Kim, J.-H., Moreira-Turcq, P., Abril, G., Hopmans, E.C., Bonnet, M.-P., Sobrinho, R.L., Sinninghe
709 Damsté, J.S., 2013. Disentangling the origins of branched tetraether lipids and crenarchaeol in the lower
710 Amazon River: Implications for GDGT-based proxies. *Limnology and Oceanography* 58, 343–353.

711 Zell, C., Kim, J.-H., Balsinha, M., Dorhout, D., Fernandes, C., Baas, M., Sinninghe Damsté, J.S., 2014.
712 Transport of branched tetraether lipids from the Tagus River basin to the coastal ocean of the Portuguese
713 margin: consequences for the interpretation of the MBT⁺/CBT paleothermometer. *Biogeosciences* 11,
714 5637–5655.

715

716 **Tables**
 717 **Table 1**
 718 Sampling station data (n.d., not determined; n.a., not available).

Site	Latitude (°N)	Longitude (°E)	Depth bsl (m) ^a	Distance SRM (km) ^b	Sampling date	Measured pH	Measured temp (°C)	POC (mg/l)	$\delta^{13}\text{C}$ (‰)
SR ^c	51.72833	107.4628	0.5	-150	06-07-2010	8.4	25.0	6.4	-27.02
SRM	52.27318	106.2547	0.5	0	20-07-2010	8.1	22.5	5.9	-28.43
S1	52.27852	106.2444	0.5	1	20-07-2010	8.1	15.0	5.1	-28.25
S3	52.29025	106.2226	0.5	3	22-07-2010	8.1	13.0	2.0	-28.46
S5	52.30313	106.1998	0.5	5	28-07-2010	8.0	11.3	1.5	-28.28
B1	52.27852	106.2444	6	1	20-07-2010	7.9	9.3	2.5	-28.29
B3	52.29025	106.2226	17	3	20-07-2010	7.9	6.6	1.1	-29.31
B5	52.30313	106.1998	25	5	28-07-2010	8.0	5.2	1.3	-29.32
Ba	51.89194	104.8233	0.5	105	01-07-2010	8.8	6.7	1.3	-24.34
MIR ^c	51.94533	100.78839	0.5	n.a.	11-07-2010	5.6	10.8	0.1	-27.85
Y1 ^c	58.00992	93.11680	2.0	n.a.	25-8-2009	n.d.	12.0	0.03	b.d.
Y2 ^c	58.13147	92.75405	3.0	1500	29-9-2009	n.d.	11.0	0.09	-30.38

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^a Below surface water level; ^b estimated downstream distance relative to Selenga River mouth, so upstream samples thus have negative values, ^c samples reported by De Jonge et al., 2014b and De Jonge et al., submitted.

745
746
747**Table 2**

Fractional abundance (% of total) and concentrations of CL brGDGTs and crenarchaeol in SPM of Selenga River and its outflow into Lake Baikal, in the Mountainous Irkut River and in the Yenisei River (b.d.l., below detection limit).

748

Site	Fractional abundance (%)														BrGDGTs (µg/g POC)	Crenarchaeol (µg/g POC)	IR	BIT	pH ^a	MST ^a (°C)	
	Ia	Ib	Ic		IIb	IIIa	IIIb	IIIc	IIa'	IIb'	IIc'	IIIa'	IIIb'	IIIc'							
SR ^b	12	5	1	0.68	4	7.5	8	0.4	b.d.l	24	7	0.4	24	1	b.d.l	4	0.2	0.68	0.93	7.5	730
SRM	13	5	1	0.71	3	7.5	7	0.3	0.1	25	8	0.3	23	1	0.3	12	0.4	0.71	0.96	7.5	20
S1	14	5	1	0.70	3	7.5	8	0.4	0.05	25	7	0.3	21	1	0.3	10	0.4	0.70	0.96	7.5	731
S3	17	7	1	0.56	5	7.1	12	b.d.l	b.d.l	18	6	b.d.l	19	b.d.l	b.d.l	1	0.4	0.56	0.71	7.1	20
S5	18	5	b.d.l	0.52	4	7.0	16	b.d.l	b.d.l	17	5	b.d.l	17	b.d.l	b.d.l	0.6	0.6	0.52	0.45	7.0	732
B1	14	6	1	0.56	4	7.2	12	0.5	b.d.l	20	6	0.3	18	1	b.d.l	6	0.9	0.56	0.84	7.2	19
B3	16	4	1	0.43	3	6.5	23	0.4	b.d.l	11	3	b.d.l	11	1	b.d.l	10	2.7	0.43	0.77	6.5	733
B5	12	5	2	0.45	6	6.9	23	1	0.3	11	4	0.4	17	1	1	2	0.8	0.45	0.63	6.9	17
Ba	10	2	b.d.l	0.44	2	7.0	26	b.d.l	b.d.l	10	1	b.d.l	32	b.d.l	b.d.l	0.1	0.2	0.44	0.22	7.0	734
MIR ^b	11	2	0.4	0.51	2	6.9	24	0.3	b.d.l	15	2	0.3	20	1	b.d.l	15	0.1	0.51	0.99	6.9	13
Y1 ^b	14	3	1	0.71	2	7.5	9	0.4	0.1	18	3	b.d.l	36	1	0.1	10	0.3	0.71	0.97	7.5	735
Y2 ^b	9	3	0.4	0.80	2	7.9	7	0.4	0.1	19	7	0.2	42	2	0.4	21	0.2	0.80	0.99	7.9	19

~~736~~757 ^areconstructed pH and MST; ^b reported by De Jonge et al., 2014b and De Jonge et al., submitted.

758

759 **Figure captions**

760 **Fig. 1.** Chemical structures of branched GDGTs (I-III) and crenarchaeol (IV). The chemical structures of the
761 hexa- and pentamethylated brGDGTs with cyclopentyl moiety(ies) IIb', IIc', IIIb' and IIIc' are tentatively assigned.

762 Fig. 2. Map of the Southern part of Lake Baikal, with the sample sites indicated. Insert **a** shows the distribution of
763 the SPM samples collected in the Selenga River outflow. Insert **b** shows, i) the extent of the watershed of Lake
764 Baikal, where all rivers that drain into Lake Baikal are indicated in orange and ii) the Southern part of the Yenisei
765 River watershed, where all rivers that drains into the Yenisei River (and the Yenisei River itself) are indicated in
766 blue.

767 **Fig. 3.** Figure plotting the environmental variables of Lake Baikal and its inflowing rivers; measured water
768 temperature ($^{\circ}\text{C}$, A), measured water pH (B), measured concentration of POC ($\mu\text{g/l}$, C) and the stable carbon
769 isotopic value of the bulk OM, $\delta^{13}\text{C}_{\text{POC}}$ (‰ , D). The shaded areas indicate riverine samples, where orange denotes
770 the Selenga River, and grey denotes the Irkut River. Surface water is indicated by round symbols, bottom water is
771 indicated with triangles. Please note that the x-axis, where the distance to the Selenga River Mouth is plotted is
772 broken, as to allow the Selenga River, the Lake Baikal sample and the Mountain River to be plotted on the same
773 axis. Selenga outflow samples, collected in the same sampling campaign, are connected.

774 **Fig. 4.** The concentration of CL brGDGTs ($\mu\text{g/g}$ POC, A) and crenarchaeol ($\mu\text{g/g}$ POC, B) and the BIT-index
775 values (C), in Lake Baikal and its inflowing rivers. The shaded areas indicate riverine samples, where orange
776 denotes the Selenga River, and grey denotes the Irkut River. Surface water is indicated by round symbols, bottom
777 water is indicated with triangles. Please note that the x-axis, where the distance to the Selenga River Mouth is
778 plotted is broken, as to allow the Selenga River, the Lake Baikal sample and the mountainous Irkut River to be
779 plotted on the same axis. Selenga outflow samples, collected in the same sampling campaign, are connected.

780 **Fig. 5.** The fractional abundance (FA), expressed as the percentage of the total of the 15 brGDGT compounds, of
781 the CL (A-F) fraction of selected samples: Selenga River Mouth (SRM, A), the surface water at 3 km distance
782 from the outflow (S3, B), the bottom water at 3 km from the outflow (B3, C), the bottom water at 5 km from the
783 outflow (B5, D), the shoreline Lake Baikal setting (Ba, E), the Mountainous Irkut River (IR, F), the Yenisei River

784 before (Y1, G) and after the Angara River inflow (Y2, H). The color of the bars refers to the structure of the
785 brGDGTs, and is referred to in the legend.

786 **Fig. 6.** PCA based on the standardized fractional abundances of the Lake Baikal brGDGTs and the inflowing
787 Selenga and Irkut Rivers. Panel A and B plot principal component 1 (PC1) against PC2, panel C and D plots PC1
788 against PC3. The scores of the brGDGT compounds are indicated in panel A and C, and the scores of the sites are
789 plotted in panels B and D. Surface waters are indicated with round symbols and connected with the corresponding
790 bottom waters (triangular symbol) with a grey line. The correlation of the measured water temperature (Temp)
791 with the PCs is plotted *a posteriori* in the ordination space (panel B), indicated with a dotted line.

792 **Fig. 7.** Truncated triplot, based on the fractional abundances of the brGDGTs Ia, IIIa and IIIa'. The sum of the
793 fractional abundances amounts up to 100%. The round symbols indicate surface water of Lake Baikal and
794 inflowing rivers. The triangles indicate Lake Baikal bottom waters, while the diamonds indicate the brGDGT
795 distribution in the Yenisei River. The full black line covers brGDGT distributions that can be explained by linear
796 mixing between the Selenga River Mouth (SRM) distribution and the Baikal outflow (Ba) distributions. The
797 dotted line covers brGDGT distributions that can be explained by linear mixing between the Selenga River Mouth
798 (SRM) distribution and the mountainous Irkut River (MIR) distributions.

799 **Fig. 8.** The isomer ratio IR [Eq. 1] (A), the reconstructed MST [Eq. 5] (B) and the reconstructed pH [Eq. 3 and 4]
800 (C) are plotted. The shaded areas indicate riverine samples, where orange denotes the Selenga River, and grey
801 denotes the mountainous Irkut River. Surface water is indicated by round symbols, bottom water is indicated with
802 triangles. Please note that at the x-axis, where the distance to the Selenga River Mouth is plotted is broken, as to
803 allow the Selenga River, the Lake Baikal sample and the Mountain River to be plotted on the same axis. Selenga
804 outflow samples, collected in the same sampling campaign, are connected.

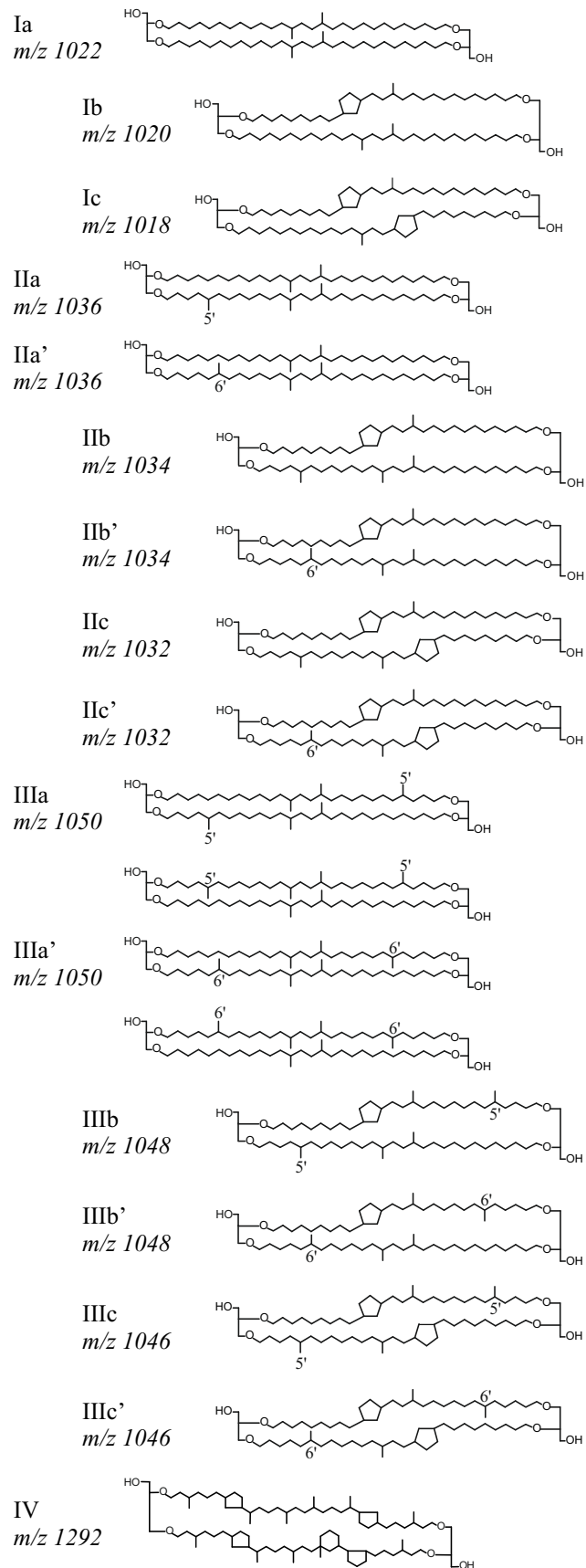


Fig. 1

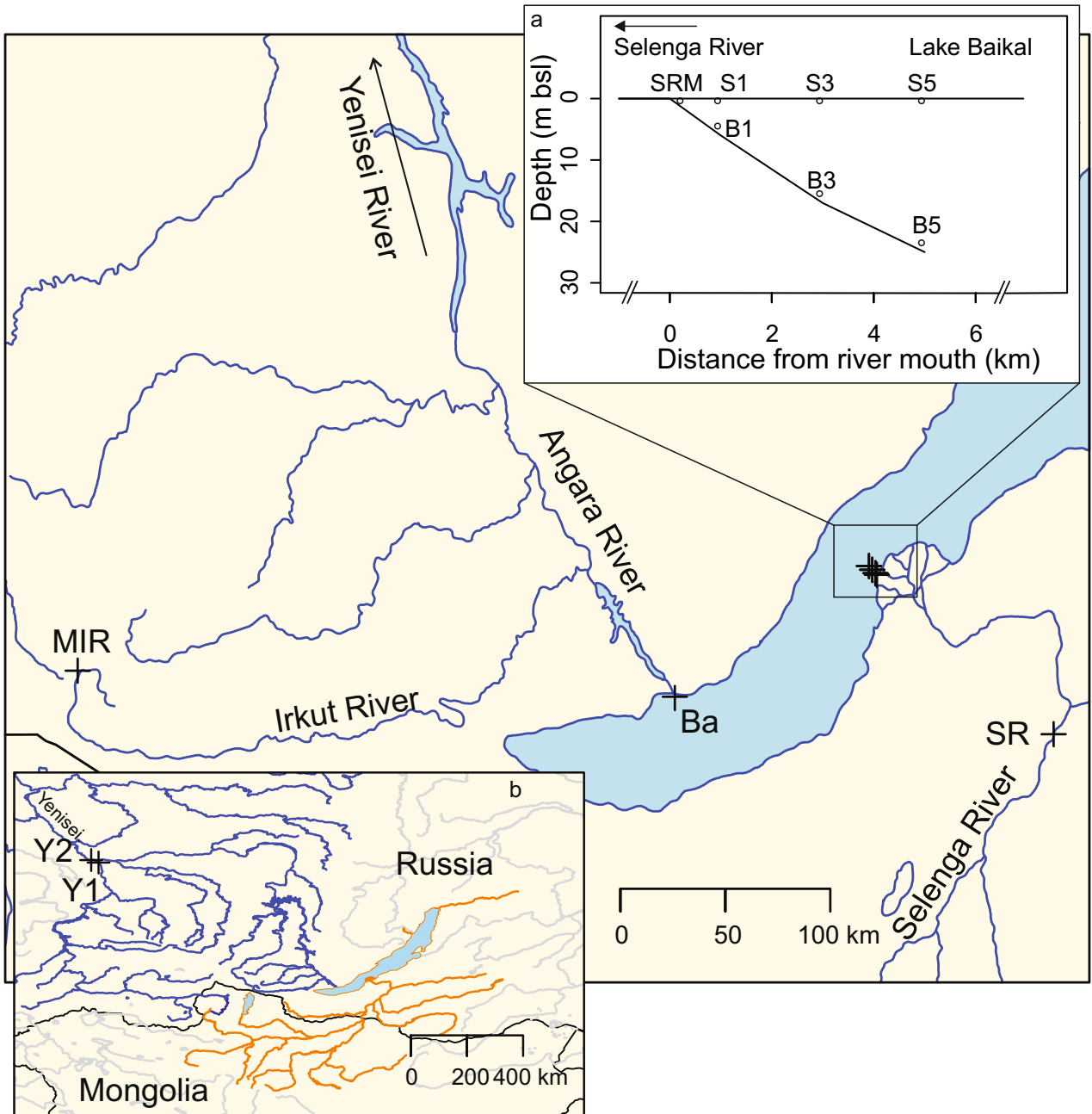


Fig. 2

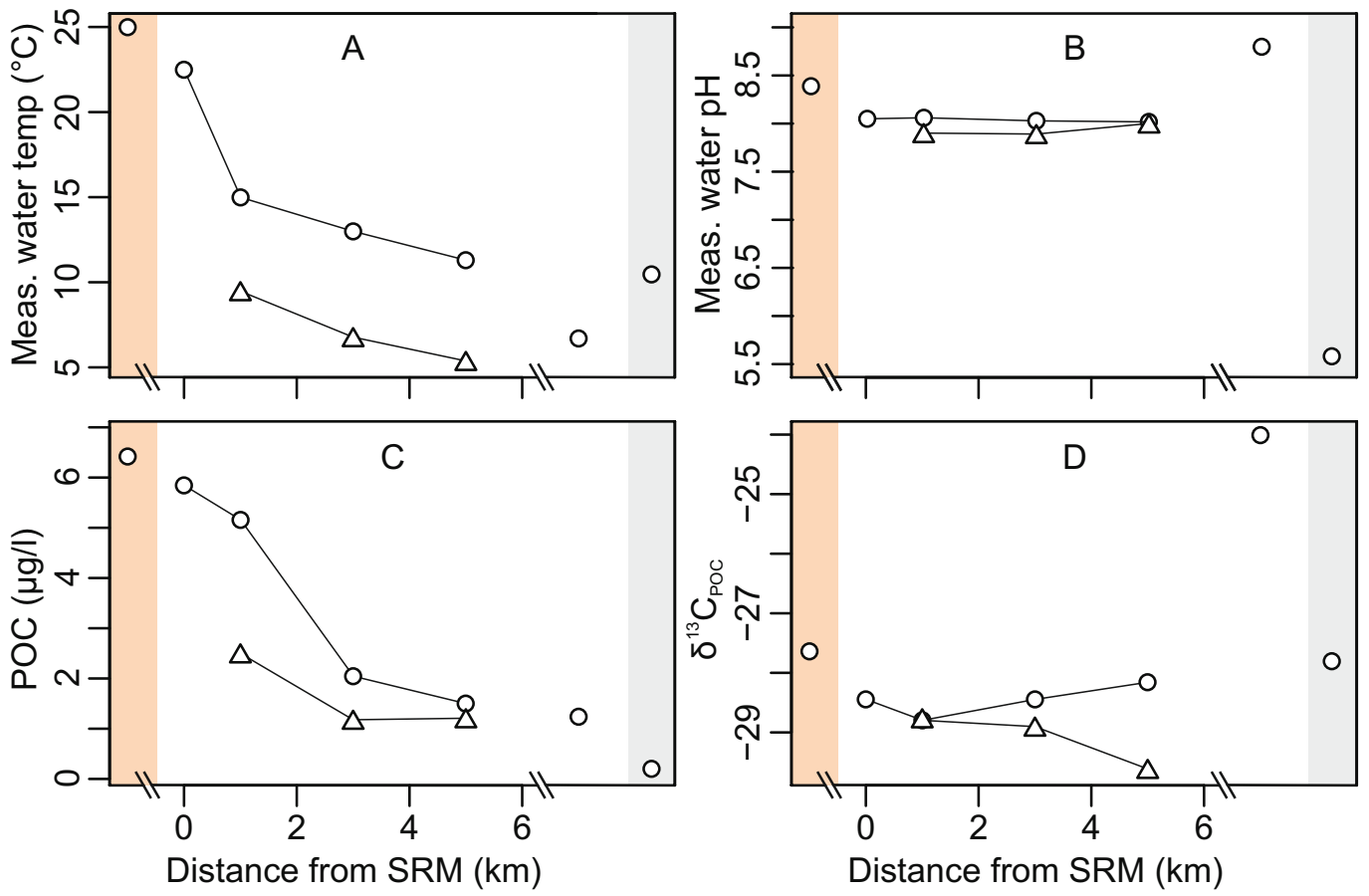


Fig. 3

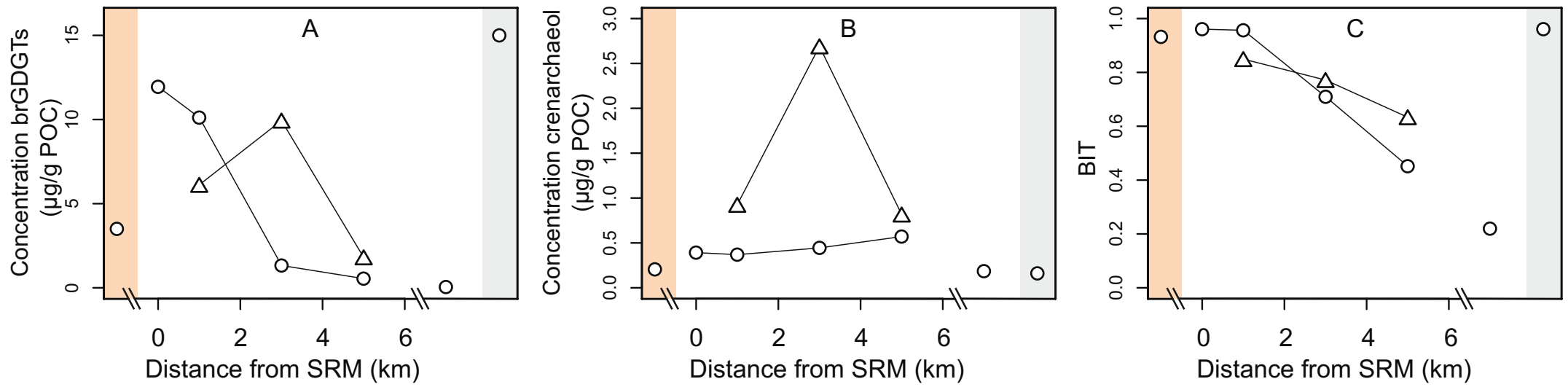


Fig. 4

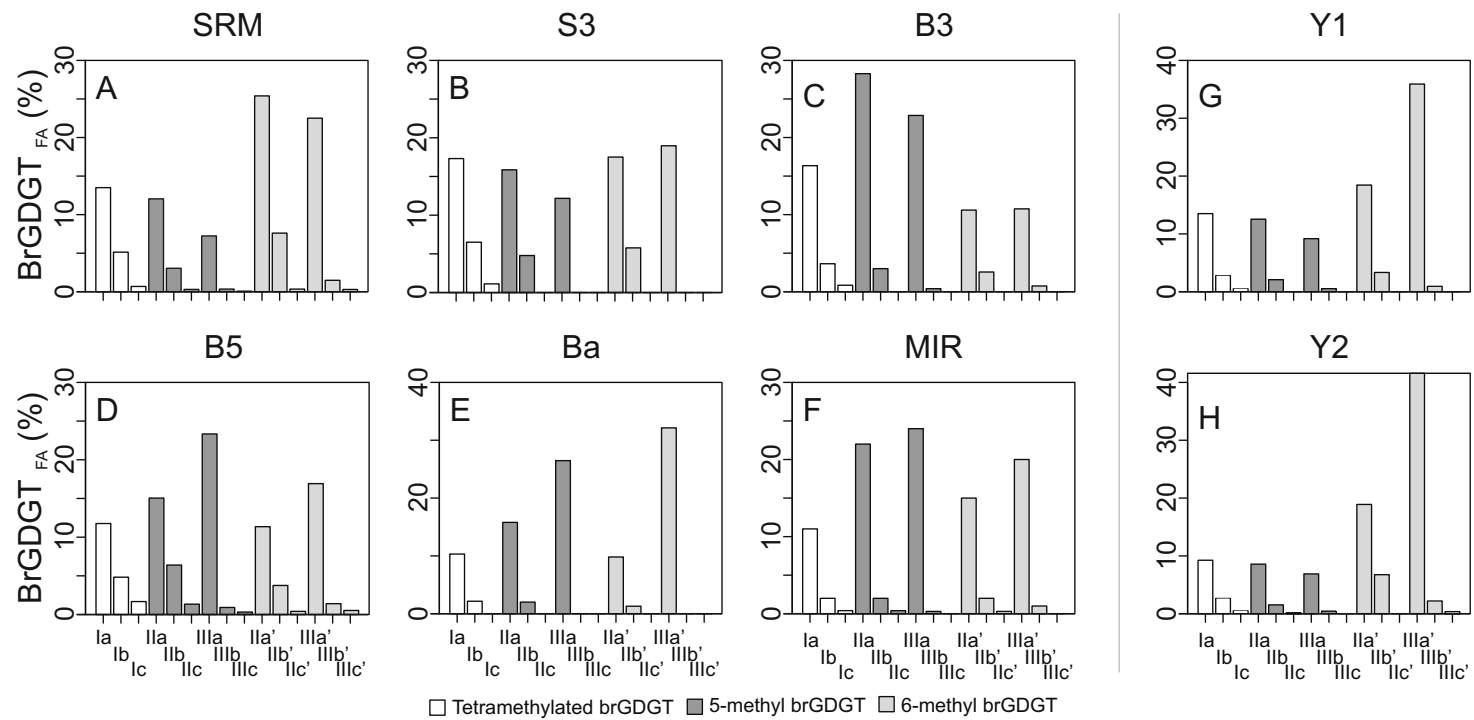


Fig. 5

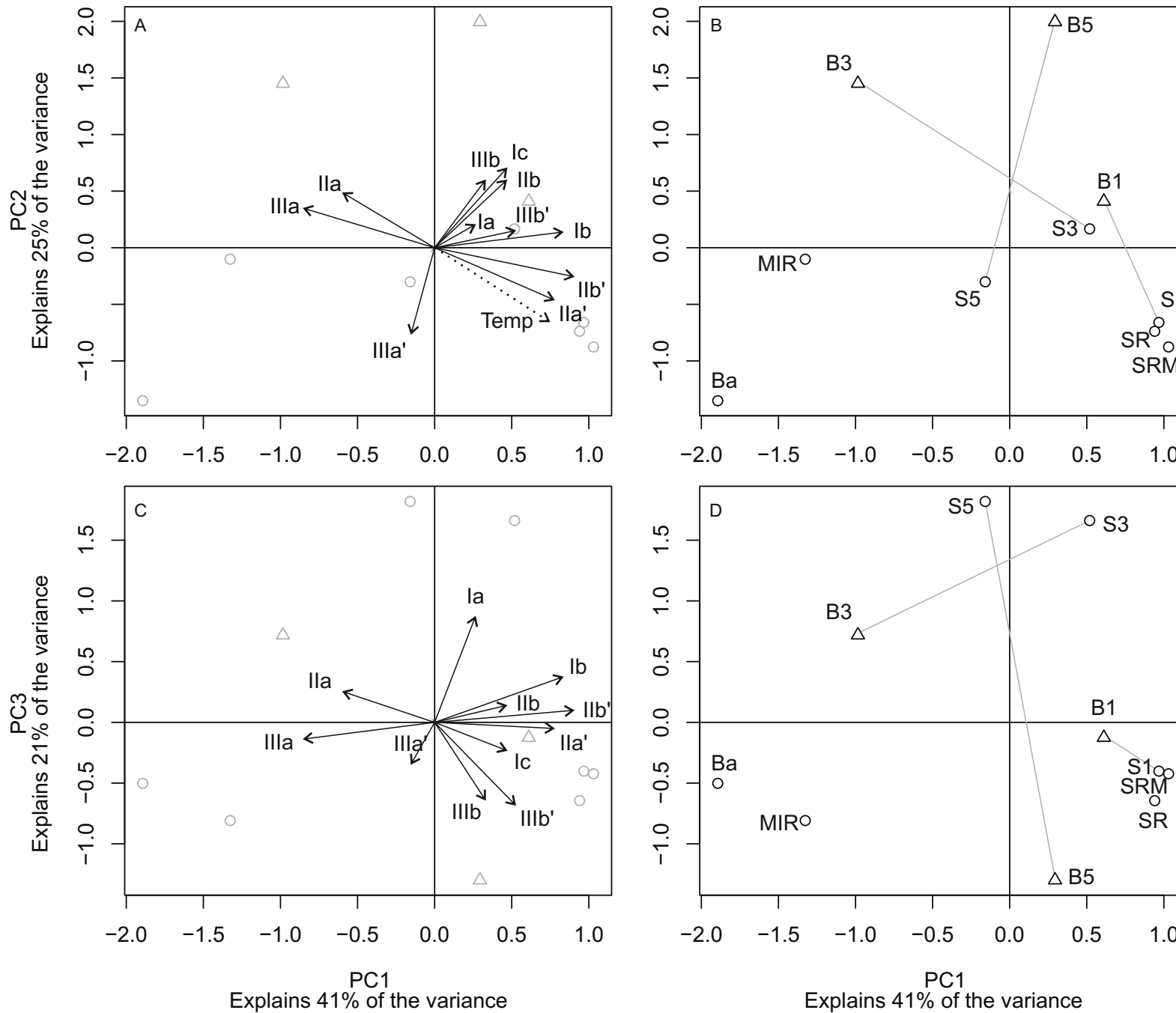


Fig. 6

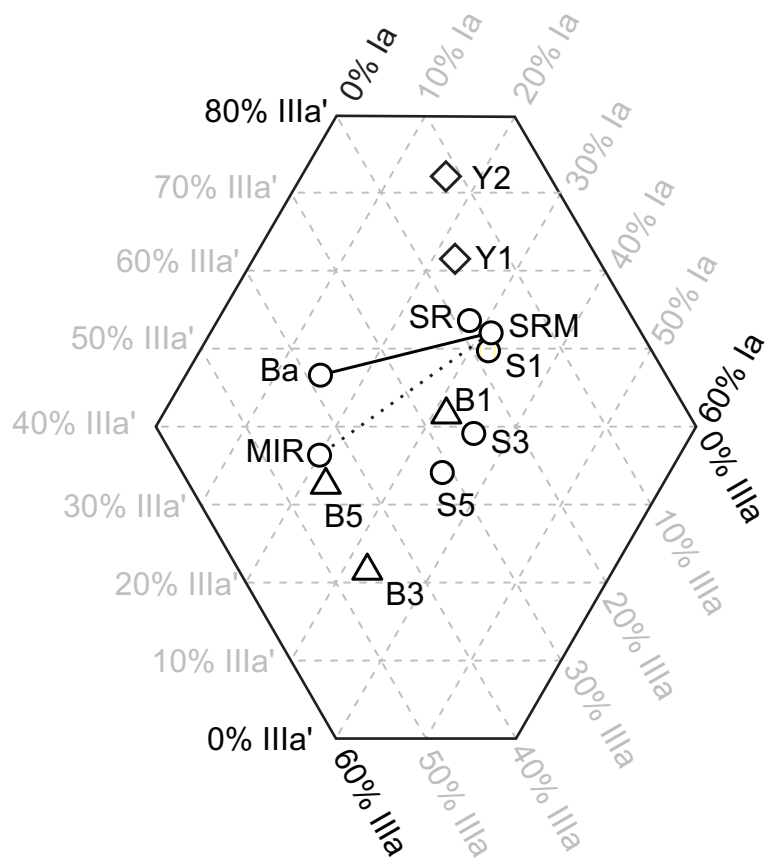


Fig. 7

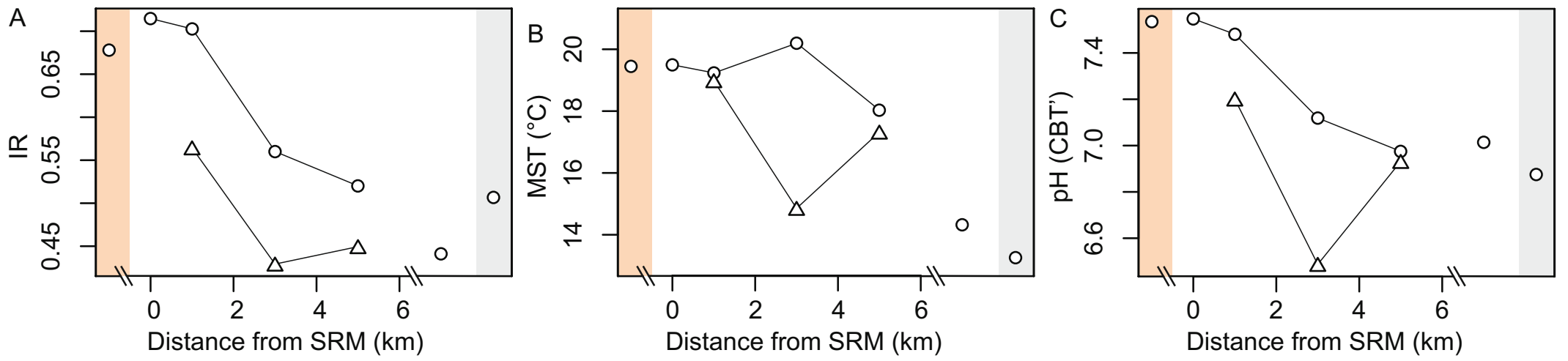


Fig. 8